

Provision of essential resources as a persistence strategy in food webs

Michael Raatz*

Department of Theoretical Biology, Max Planck Institute for Evolutionary Biology, Plön, Germany

Pairwise interactions in food webs, including those between predator and prey are often modulated by a third species. Such higher-order interactions are important structural components of natural food webs that can increase the stability of communities against perturbations and ensure continued ecosystem functioning. Particularly the flux of rare organic and inorganic compounds that are essential to species in the community can create higher-order interactions. Even though many such compounds exist, their effect on structuring communities is little understood. In this study, I perform invasion analyses on a general food web model that depicts apparent and exploitative competition. Introducing the provision of essential resources by a prey species to either its competitor or its predator as a higher-order interaction, I find that this mechanism can ensure the focal prey's persistence. Larger dietary essentiality, i.e. a stronger dependence of the predator or the competitor on the essential resource can increase the invasion growth rate of the focal prey to positive values, thus promoting its persistence when it would go extinct for low essentiality. This research shows that essential resources and the higher-order interactions created by them should be considered in community ecology.

Keywords: Higher-order interaction, Food webs, Food quality, Coexistence, Invasibility

*mraatz@evolbio.mpg.de

22 Introduction

23 Growth, reproduction and survival of organisms can be limited by organic and inorganic compounds,
24 which often are not present in the organism's diet in favourable concentrations or ratios. Particularly
25 consumers at the plant-herbivore interface are often affected by dietary mismatches (Elser et al.,
26 1996; Gaedke et al., 2002; Wacker & Martin-Creuzburg, 2012; Urabe et al., 2018). This motivated
27 considering besides food quantity also the quality of food when investigating performance measures of
28 aquatic consumers (Andersen et al., 2004; Anderson & Hessen, 2005; Wacker & Martin-Creuzburg,
29 2012; Guo et al., 2016; Schälicke et al., 2019; Koussoroplis et al., 2019), terrestrial herbivores
30 (Douglas, 2015; Eberl et al., 2020) and pollinators (Filipiak et al., 2017). The scarcity of resources
31 that are essential for growth and reproduction but cannot be easily acquired from the environment
32 can constrain the flow of matter and energy between trophic levels. Therefore, dietary limitations
33 induced by essential resources can have important effects on population and community dynamics
34 (Muller et al., 2001; Gaedke et al., 2002; Schade et al., 2003; Stiefs et al., 2010; Iwabuchi & Urabe,
35 2012; Singer et al., 2012; Raatz et al., 2017; Burian et al., 2020).

36 Dietary dependencies also regularly exist within the same trophic level, where uptake of essential
37 resources occurs from the environment, such as within the microbial loop when bacteria consume
38 dissolved organic carbon from phytoplankton exudates (Azam et al., 1983; Pomeroy et al., 2007) or
39 during the exchange of essential nutrients and metabolites between bacteria and microalgae (Soria-
40 Dengg et al., 2001; Croft et al., 2005; Kazamia et al., 2012; Suleiman et al., 2016; D'Souza et al.,
41 2018; Oña & Kost, 2022). Understanding the mechanisms and effects of such dependencies is crucial
42 for biodiversity research given for example the importance of microalgae such as diatoms for aquatic
43 ecosystems and global carbon dynamics (Croft et al., 2005; Koedooder et al., 2019), but also for
44 medical fields like human microbiome research (Herren, 2020) and antibiotic resistance research
45 (Adamowicz et al., 2018). Taken together, dietary mismatches and dependencies of organisms from

46 the same or different trophic levels are crucial determinants for the structure of their communities.

47 Mechanistically, community structure is determined by direct interactions within pairs of species

48 or by indirect interactions across multiple species from the same or different trophic levels, e.g.

49 through trophic cascades or apparent competition. Additionally to direct and indirect interactions,

50 higher-order interactions, here defined as the density of a third species affecting the interaction of

51 two other species (sensu [Billick & Case, 1994](#)), were found to potentially structure communities.

52 The effects of higher-order interactions include stabilizing population dynamics ([Grilli et al., 2017](#)),

53 increasing robustness against perturbation ([Terry et al., 2019](#); [Gibbs et al., 2023](#)), determining

54 fitness of competitors ([Mayfield & Stouffer, 2017](#)) and affecting biodiversity-ecosystem-functioning

55 relationships ([Miele et al., 2019](#)). Examples for higher-order interactions include trait-mediated

56 effects such as a predator affecting the foraging rate of its prey or the prey's predation risk from

57 other predators, and environment-mediated effects such as one species providing refuge to another

58 species ([Wootton, 2002](#); [Miele et al., 2019](#)).

59 In this paper, I will investigate another, so far unrecognized mechanism for creating higher-order

60 interactions that arises from the provisioning of essential resources. In the presence of dietary

61 mismatches one species, from here on referred to as the focal species, may provide resources that

62 are essential to other community members. Such interactions are possible both towards members

63 of the same trophic level, such as competitors, or towards members of different trophic levels, e.g.

64 shared predators that prey on multiple species. For example, a higher-order interaction within the

65 same trophic level is created when a competitor is co-limited by two resources but can only obtain one

66 of those two resources from its environment and relies on another prey (the focal prey) to provide the

67 other co-limiting resource. This provision may occur for example by leakage of common goods ([Gore](#)

68 [et al., 2009](#)) or carbon exudation in otherwise carbon-limited environments ([Bratbak & Thingstad,](#)

69 [1985](#); [Raatz et al., 2018](#)). A higher-order interaction between different trophic levels can arise when

70 a predator obtains energy from multiple prey species but only the focal prey may provide additional,
71 essential resources, e.g. vitamins or polyunsaturated fatty acids (Wacker & Martin-Creuzburg, 2012).
72 Excess essential resources provided by the focal prey may then be used to efficiently convert other
73 low-quality prey into predator biomass (Raatz et al., 2017).
74 In these two cases the provision of essential resources by the focal prey creates a higher-order inter-
75 action that manifests as an interaction modification (sensu Terry et al., 2019) that regulates the flow
76 of matter to the competitor or predator compartment in these communities, respectively. Regulating
77 such fluxes has the potential to affect the biomass distribution in the community, ultimately deter-
78 mining the persistence of individual species. If such higher-order interactions increase the persistence
79 of the focal prey they pose as an example for a niche-improving form of niche construction and they
80 may thus even be adaptive (Kylafis & Loreau, 2008, 2011; Laland et al., 2016). Consequently, in
81 this paper, I will establish the provision of essential resources in a community as a mechanism driving
82 higher-order interactions that may increase the persistence of the focal prey species and prevent its
83 extinction either from predation or competition.

84 **Methods**

85 Investigating persistence of a focal species typically employs invasion analysis, which determines the
86 net growth rate of that species in the remaining resident community when it is rare (and assumed
87 to be invading) (MacArthur & Levins, 1967; Chesson, 1994; Ellner et al., 2019). If the focal
88 prey provides the essential resources to some components of the community, being rare equates
89 to switching off the higher-order interaction. Invasion analysis is therefore the perfect tool for
90 determining the effect of essential resources both on the resident community and the persistence of
91 the focal prey. Accordingly, I will investigate the invasion growth rate of the focal prey species X_1 in

92 a community that contains an abiotic resource R , a competing species X_2 and a shared predator Y
 93 (Eqn. 1, Fig. 1), to incorporate essential resource provision in food webs. I assume a chemostat-type
 94 model in which the abiotic resource R is provided at a constant rate $R_0 \delta$ and all entities experience
 95 the same dilution, see Tab. 1 for parameter definitions and values. The two prey species X_1 and X_2
 96 take up the abiotic resource R at some rate $r u(R)$, where $u(R)$ defines the functional form of prey
 97 resource uptake. Both prey species are consumed by the predator following a functional response
 98 $f(X_1, X_2)$. I assume that the two prey species differ only in their attack probability (sensu Ehrlich
 99 & Gaedke, 2018) by a factor p and their maximum growth rate by a factor ϕ . For example, $p < 1$
 100 and $\phi < 1$ implements a growth-defense trade-off (Fig. 2b).

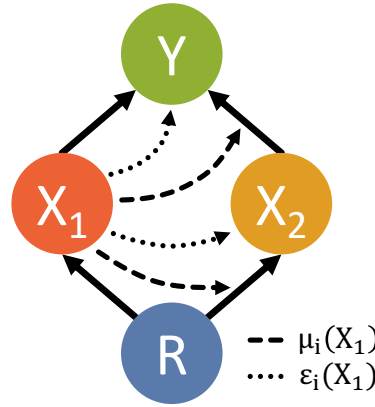


Figure 1 Food web structure. The model equations (Eq. 1) describe a diamond-shaped food web module. Solid arrows depict flows of matter due to resource or prey uptake. Dashed arrows show the interaction modification $\mu_i(X_1)$ of the uptake rates that are caused by the provision of essential resources by the focal prey. The other potential higher-order interaction from essential resource provision $\varepsilon_i(X_1)$ affects the conversion efficiency of the competitor or the predator and is depicted by dotted arrows. Only one of these higher-order interaction is investigated at a time.

$$\begin{aligned}
 \frac{dR}{dt} &= (R_0 - R) \delta - r u(R) X_1 - \mu_X(X_1) \phi r u(R) X_2 \\
 \frac{dX_1}{dt} &= r u(R) X_1 - f(X_1, X_2) X_1 Y - \delta X_1 \\
 \frac{dX_2}{dt} &= \varepsilon_X(X_1) \mu_X(X_1) \phi r u(R) X_2 - \mu_Y(X_1) p f(X_1, X_2) X_2 Y - \delta X_2 \\
 \frac{dY}{dt} &= f(X_1, X_2) X_1 Y + \varepsilon_Y(X_1) \mu_Y(X_1) p f(X_1, X_2) X_2 Y - \delta Y
 \end{aligned} \tag{1}$$

101 Throughout this paper, I use a Monod-type term to indicate resource limitation of the prey

$$u(R) = \frac{R}{K + R} \quad (2)$$

102 and a Holling Type-2 functional response for multiple prey species to describe the predation rate by
103 an individual predator:

$$f(X_1, X_2) = g \frac{1}{H + X_1 + p X_2} \quad (3)$$

104 I introduce the higher-order interactions due to essential resource provision as interaction modifica-
105 tions $\mu_i(X_1)$ and $\varepsilon_i(X_1)$ driven by the density of the focal prey species (Arditi et al., 2005). For
106 generality, I include all possible options where these modifications affect the uptake rates of abiotic
107 resources or prey, or the efficiency at which new biomass is produced, respectively. Accordingly,
108 $\mu_X(X_1)$ defines how an increasing density of the focal prey increases the resource uptake rate of the
109 competing prey and $\varepsilon_X(X_1)$ gives the conversion efficiency of those resources into new competitor
110 biomass depending on the density of the focal prey. The same logic translates to $\mu_Y(X_1)$ and $\varepsilon_Y(X_1)$
111 for the predator. i.e. the provision of essential resources may increase the predation rate on the
112 competitor, for example by alleviating a predator dispreference for the competitor. Similarly, the
113 predator conversion efficiency for competitor biomass may increase due to the provision of essential
114 resources by the focal prey. Note that in this model, I am investigating only the provision of es-
115 sential resources, thus assuming that the focal prey itself always contains the essential resources.
116 Predator consumption and conversion of focal prey biomass is thus kept constant. I assume that the
117 modification functions $\mu_i(X_1)$ and $\varepsilon_i(X_1)$ monotonically increase with focal prey density, eventually
118 approaching unity for large focal prey densities, as here their effect should vanish, as the essential
119 resource should be abundantly present and thus non-limiting.

$$\lim_{X_1 \rightarrow \infty} \mu_i(X_1) = 1$$

$$\lim_{X_1 \rightarrow \infty} \varepsilon_i(X_1) = 1$$

120 The interaction modifications pose an implicit way of representing the temporal dynamics of produc-
 121 tion, distribution, stability, uptake and usage of the essential resource molecules and thus avoid the
 122 difficulties involved in modelling these processes explicitly, but explicit approaches also exist (Sun
 123 et al., 2019; Hammarlund et al., 2019).

124 I define essentiality e as the relative reduction of uptake rates or conversion efficiencies in the absence
 125 of the focal prey compared to when it's abundantly present and neither the uptake rates nor the
 126 conversion efficiencies are reduced. For the uptake rate modifications $\mu_i(X_1)$ this results in

$$e = \frac{\lim_{X_1 \rightarrow \infty} \mu_i(X_1) - \mu_i(0)}{\lim_{X_1 \rightarrow \infty} \mu_i(X_1)} = 1 - \mu_i(0) \quad (4)$$

127 A high essentiality thus implies a strong reduction in the uptake rates when the focal prey is absent.

128 Similarly, if the higher-order interaction is incorporated into the conversion efficiencies I define

$$e = 1 - \varepsilon_i(0) \quad (5)$$

129 Note that for the sake of simplicity I investigate only the effect of one higher-order interaction at a
 130 time, i.e. only one interaction modification will be dependent on the focal prey density, keeping the
 131 other three constant at unity.

132 To determine the effect of essential resource provision on persistence of the focal prey I employ
 133 invasion analysis and study whether the focal prey can re-invade the resident community once it has
 134 gone extinct. This is ensured by a positive invasion growth rate which is defined as the average
 135 per-capita growth rate when rare (Ellner et al., 2019). Specifically, the invasion growth rate of the
 136 focal prey in my model is

$$\left\langle \frac{1}{X_1} \frac{dX_1}{dt} \Big|_{X_1=0} \right\rangle = \langle r u(R) - f(0, X_2) Y - \delta \rangle \quad (6)$$

137 where the angle brackets denote the temporal average. If the resident community's attractor is a
 138 limit cycle, the temporal average can be obtained numerically from one period of such cycles (Ehrlich

Table 1 Reference parameter set. Resource concentrations and organism abundances or densities are assumed to be normalized appropriately such that their units become unity. Deviations from the reference parameter values are reported where applicable. For an illustration of the different food web scenarios see Fig. 2.

Parameter	Biological meaning	Value			
R_0	Input concentration of abiotic resource	1			
δ	Chemostat dilution rate	1 time unit ⁻¹			
r	Prey's uptake rate	2 time unit ⁻¹			
K	Prey's half-saturation constant	0.1			
g	Predator's consumption rate	1.5 time unit ⁻¹			
H	Predator's half-saturation constant	0.1			
		Food web scenarios			
		I	II	III	IV
ϕ	Relative competitiveness of the competitor	1.05	0.95	0.95	1.05
p	Relative vulnerability of the competitor to predation	0.8	0.8	1.2	1.2

139 [et al., 2017](#)). As at least some of the parameter combinations investigated in this paper result in
140 limit cycles I used this numerical approach throughout and confirmed the results with the analytically
141 computable solutions for those cases where the resident community was in a steady state. For numer-
142 ically determining the invasion growth rate of the focal prey, the resident community dynamics were
143 numerically integrated for 2000 time units until they reached their attractor. Convergence was deter-
144 mined visually [and verified by ensuring that the absolute values of the slope of the linear regression](#)
145 [on the predator abundance, as well as the slope of its moving variance was below 10⁻³ per time](#)
146 [unit](#). The period length was determined as in [Raatz et al. \(2019\)](#) by determining the average time
147 spans between predator maxima during the last 200 time units using the FindMaximum algorithm
148 in Mathematica. The average of the invasion growth rates for each time step during one period was
149 computed. All computations were performed in Mathematica ([Wolfram Research, Inc., 2019](#)) and
150 can be re-run using the provided Mathematica notebooks (DOI [10.5281/zenodo.8093740.7575588](https://doi.org/10.5281/zenodo.8093740.7575588)).
151 The analytical solutions are lengthy and can also be found in the notebook and the corresponding
152 pdf exports.
153 For evaluating the state of the resident community as well as the invasion growth rate of the focal

154 prey, the interaction modifications $\mu_i(X_1)$ and $\varepsilon_i(X_1)$ reduce to $\mu_i(0) = \mu_i^0$ and $\varepsilon_i(0) = \varepsilon_i^0$, which I
 155 can relate to the essentiality e via Eqs. 4 and 5. Therefore, I obtain direct relationships between the
 156 essentiality of the resource that is provided by the focal prey and its invasion growth rate.
 157 Notably, the invasion analysis does not require a specific choice of the functional form of the interac-
 158 tion modifications. Only those numerical integrations where the focal prey is not set to zero require
 159 a particular definition. In those cases I use the following functions that monotonically increase and
 160 saturate at unity for large X_1 .

$$\begin{aligned}\mu_i(X_1) &= \mu_i^0 + (1 - \mu_i^0) \frac{X_1}{0.5 R_0 + X_1} \\ \varepsilon_i(X_1) &= \varepsilon_i^0 + (1 - \varepsilon_i^0) \frac{X_1}{0.5 R_0 + X_1}\end{aligned}\tag{7}$$

161 To investigate under which conditions the provision of essential resources can ensure persistence
 162 I will focus on four food web scenarios that account for the non-trivial coexistence outcomes in
 163 the diamond-shaped food web module. In the first food web scenario the focal prey is the inferior
 164 competitor for resource R and is more vulnerable to predation than its competitor, which would
 165 imply extinction of the focal prey without essentiality (Fig. 2a, see black arrows). In the second food
 166 web scenario the focal prey is again more vulnerable to predation but now the superior competitor
 167 for resource R , which allows for predator-mediated coexistence for a subset of the parameter space,
 168 but focal prey extinction otherwise (Abrams, 1999; Jones & Ellner, 2007) (Fig. 2b). The third
 169 and fourth food web scenarios are the mirror images of scenarios one and two (Fig. 2c and d).
 170 Complementing these scenarios, I will scan the parameter space of vulnerability to predation p and
 171 resource competitiveness ϕ of the competitor relative to the focal prey.

172 A priori one would expect that essentiality that limits the growth and competitiveness of the com-
 173 petitor should favour the persistence of the focal prey. Further, I hypothesize that within predator-
 174 mediated coexistence an increasing essentiality should make the focal prey more indispensable to
 175 the community and therefore increase its invasion growth rate, possibly even eventually fulfilling the

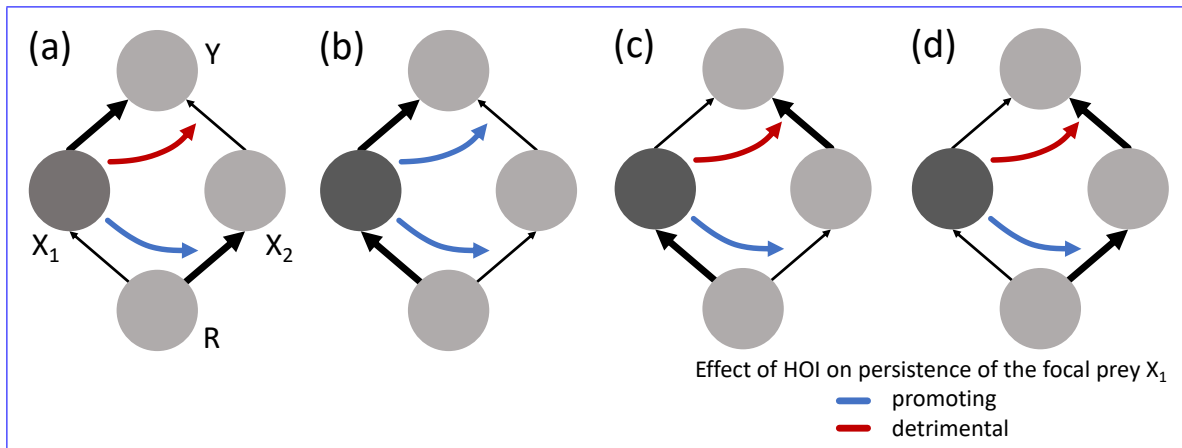


Figure 2 The effect of higher-order interactions depends on the food web scenario. In the first food web scenario (a), the focal prey X_1 is more vulnerable to predation and less competitive than its competitor X_2 , whereas it is more vulnerable to predation but also more competitive in the second food web scenario (b). The third (c) and fourth (d) food web scenarios are mirror images of the first and second food web scenario. Essentiality-mediated higher-order interactions that limit the growth of the competitor should favour persistence of the focal prey X_1 (blue curved arrows). Essentiality should promote persistence of the focal prey in food webs that permit predator-mediated coexistence (blue curved arrow in (b)), but likely is detrimental otherwise (red curved arrows) as it can render the competitor effectively less vulnerable to predation than the focal prey.

176 invasion criterion

$$\left\langle \frac{1}{X_1} \frac{dX_1}{dt} \Big|_{X_1=0} \right\rangle > 0$$

177 that would prove an ensured persistence of the focal prey.

178 Results

179 Scanning the parameter space of vulnerability to predation p and resource competitiveness ϕ of
 180 the competitor relative to the focal prey provides an overview of the effects of essentiality on the
 181 persistence of the focal prey (Fig. 3). Comparing the invasion growth rates at vanishing and complete
 182 essentiality, I find that depending on these parameters, and thus the respective food web scenario,
 183 essentiality-mediated higher-order interactions can promote but also counter-act the persistence
 184 of the focal prey, or have no effect as the focal prey persists or goes extinct irrespective of its
 185 essentiality. Analysing the four food web scenarios in more detail provides a detailed understanding

of the mechanisms behind these patterns.

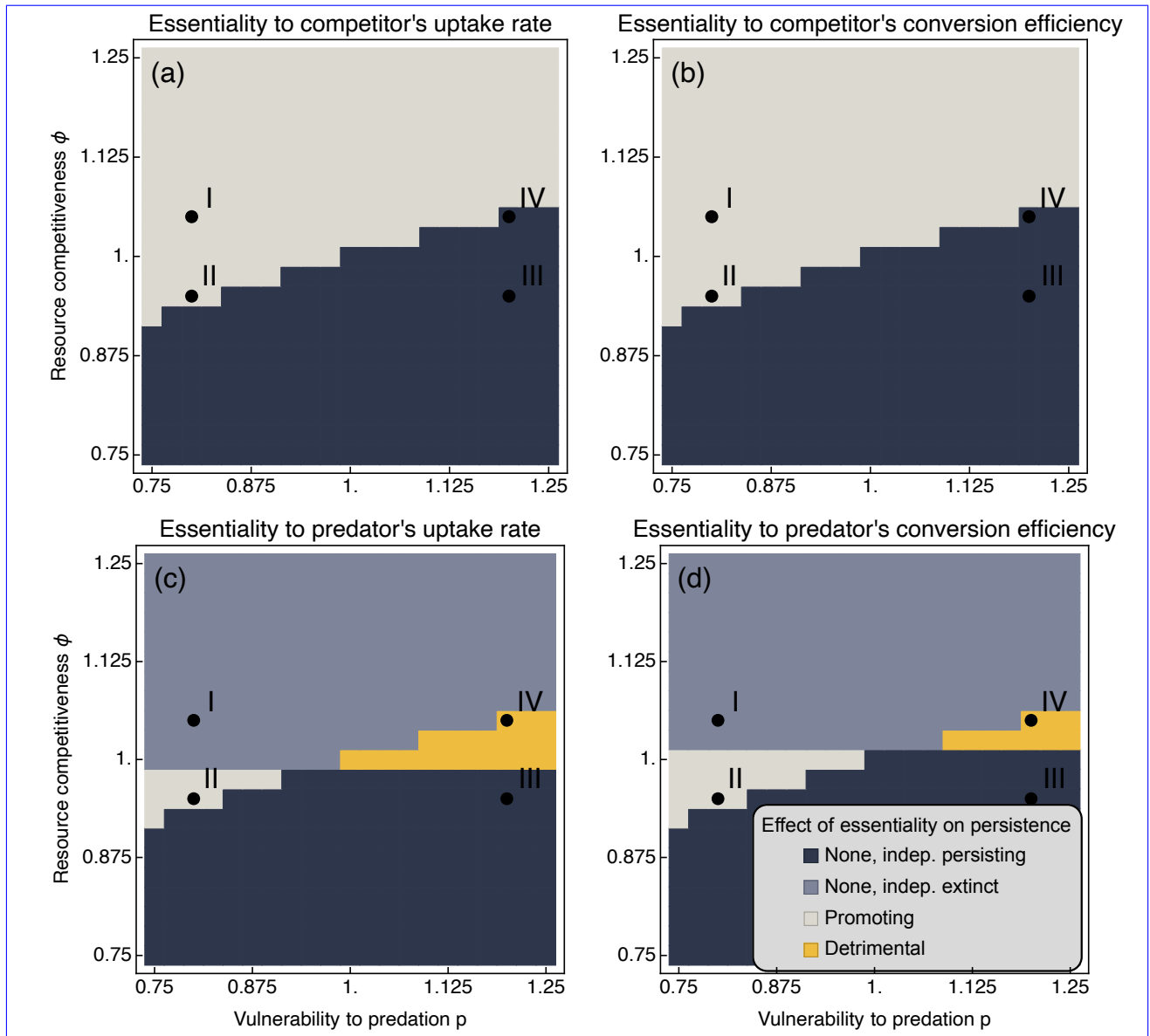


Figure 3 Effect of essentiality on the persistence of the focal prey Comparing the invasion growth rates of the focal prey for essentialities of $e = 0$ and $e = 1$ allows to classify the effects of essentiality on persistence for the four higher-order interactions indicated in Fig. 1. If the invasion growth rate for vanishing and complete essentiality are both positive then the focal prey persists independent of its essentiality (dark blue region). Vice versa, if both invasion growth rates are negative the focal prey goes extinct independent of its essentiality (light blue region). Sign changes from negative to positive for increasing essentiality indicate a persistence-promoting effect of essentiality (light-grey region), whereas sign changes from positive to negative depict a detrimental effect of essentiality on persistence of the focal prey (yellow region). The parameters of the four food web scenarios of Fig. 2 are indicated by Roman numerals.

186

187 In the first food web scenario (Fig. 2a) the focal prey does not persist for vanishing essentiality,

188 as indicated by a negative invasion growth rate. However, increasing essentiality when the higher-

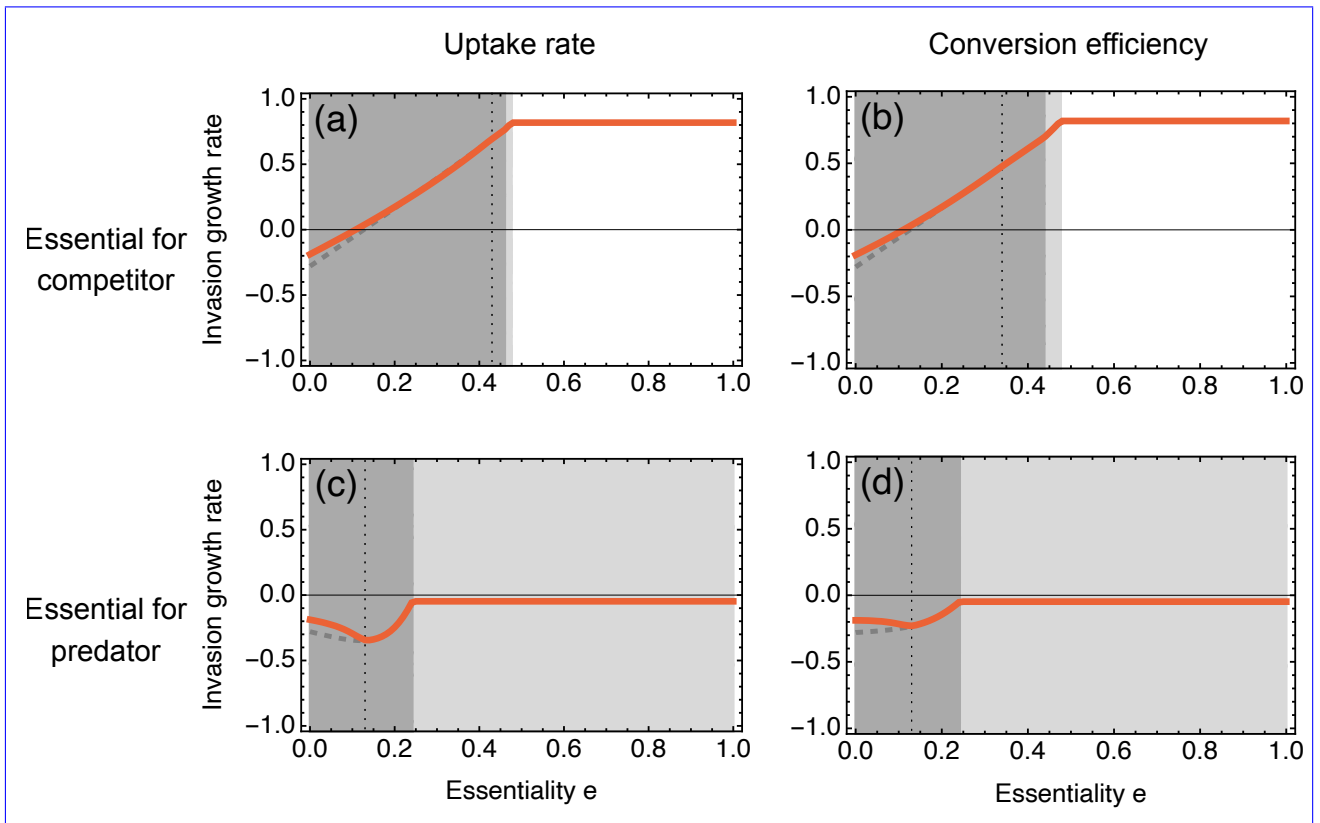


Figure 4 Invasion growth rate of the focal prey for the first food web scenario (Fig. 2a). Here, the focal prey is more vulnerable to predation and competitively inferior to the competitor. Essential resource provisioning affects (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. The grey shading indicates the states of the resident community (Fig. A1). For dark shading both predator and competitor coexist, for light-grey shading only the competitor persists and for no shading only the resource remains. The analytically computed invasion growth rate (dashed line) deviates from the numerical observation (full line) once the dynamics become cyclic. The vertical dotted line marks the bifurcation point.

189 order interaction affects the resource uptake rate or conversion efficiency of the competitor turns
 190 the invasion growth rate positive (Fig. 4a,b) and thus ensures the persistence of the focal prey
 191 (Fig. 5). This includes a drastic shift in the resident community shortly beyond $e = 0.4$ where
 192 first the predator and then the prey go extinct (Fig. A1a,b). An essentiality of $e = 0.4$ implies
 193 that the resource uptake rate or the conversion efficiency of the competitor are reduced to 60% in
 194 the absence of the focal prey. In my model formulation this implies that the competitor cannot
 195 sustain the predator further which, in the absence of the focal prey, therefore goes extinct. A slight
 196 additional reduction hinders the competitor from outgrowing dilution and thus drives it to extinction

197 as well. In this food web scenario, higher-order interactions that target the uptake rate or the
 198 conversion efficiency of the predator do not benefit the persistence of the focal prey (Fig. 4c,d) due
 199 to unfavourable trait combinations. As the focal prey is the inferior competitor for the resource R
 200 and also more vulnerable to predation it can persist neither in the absence nor in the presence of the
 201 predator. Supporting the predator by providing essential resources harms the focal prey more than
 202 the competitor. For the predator, a larger dependence on the focal prey is also disadvantageous as
 203 this decreases its uptake rate and conversion efficiency, and results in extinction at approximately
 204 $e = 0.25$ (Fig. A1c,d).

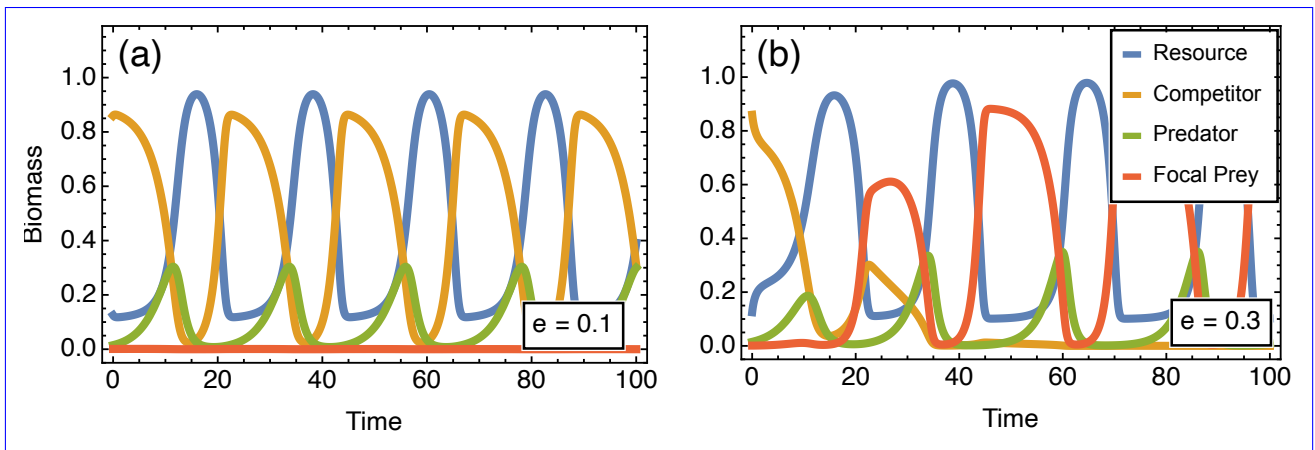


Figure 5 Population dynamics for the first food web scenario when the higher-order interaction targets the resource uptake rate of the competitor (Fig. 4a). (a) For small essentialities the invasion rate of the focal prey is negative and it thus cannot invade. (b) A larger essentiality ensures the persistence of the focal prey. To obtain these dynamics, I chose the interaction modification according to Eq. 7, integrated the resident community to its stable state and then introduced the focal prey at an initial biomass of $X_{1,0} = 10^{-3}$.

205 In the second food web scenario, the focal prey is still more vulnerable to predation than its competitor but now it is also the superior competitor for the resource R (Fig. 3). While the invasion
 206 analysis outcomes are similar to the first food web scenario for higher-order interactions targeting
 207 the competitor's uptake rate or conversion efficiency (Fig. 6a,b), the trait combinations now allow
 208 for positive invasion growth rates also when the higher-order interaction targets the predator's uptake rate or conversion efficiency (Fig. 6c,d). Therefore, increasing essentiality can promote the
 209
 210

211 persistence of the focal prey for intermediate to large essentiality in this food web scenario. This
212 persistence-promoting effect of essentiality appears in a parameter range of predator-mediated coex-
213 istence of prey (Fig. 7). Here, the predator goes extinct in the resident community as the competing
214 prey alone does not sustain the predator given the reduction in uptake rate or conversion efficiency
215 for large essentiality of the focal prey (Fig. A2). In the absence of the predator the focal prey ben-
216 efits from its higher competitiveness for the resource R and thus persists. Once it invades it may
217 additionally sustain the predator (Fig. 7b). Conditional on the presence or absence of the predator
218 when the focal prey invades two community states are therefore possible. Without the predator the
219 focal prey out-competes the competitor which thus goes extinct (Fig. 7b, solid lines). If the predator
220 is present or is re-introduced it however mediates coexistence of the focal prey and the competitor
221 (Fig. 7b, dashed lines). This shows that providing essential resources can affect not only the focal
222 prey itself, but also the whole community structure.

223 In the third food web scenario, the focal prey persists independent of essentiality as indicated by a
224 positive invasion growth rate for all possible types of essentiality-mediated higher order interactions
225 (Fig. 3 and Fig. A3). If essentiality affects the competitor the focal prey's invasion growth rate
226 increases further. If, however, essentiality causes limitations for the predator the invasion growth
227 rate tends to decrease for larger essentiality (albeit not turning negative) as this effectively reduces
228 the energy flow from the competitor to the predator and thus eventually renders the competitor less
229 vulnerable to predation than the focal prey.

230 Similarly, higher-order interactions affecting the competitor increase the invasion growth rate of the
231 focal prey with higher essentiality in the fourth food web scenario. If the resource competitiveness
232 of the competitor is only slightly exceeding the resource competitiveness of the focal prey the in-
233 vasion growth rate of the focal prey is positive even for zero essentiality and only increases further
234 for higher essentiality (Fig. A4). For higher competitiveness of the competitor the invasion growth

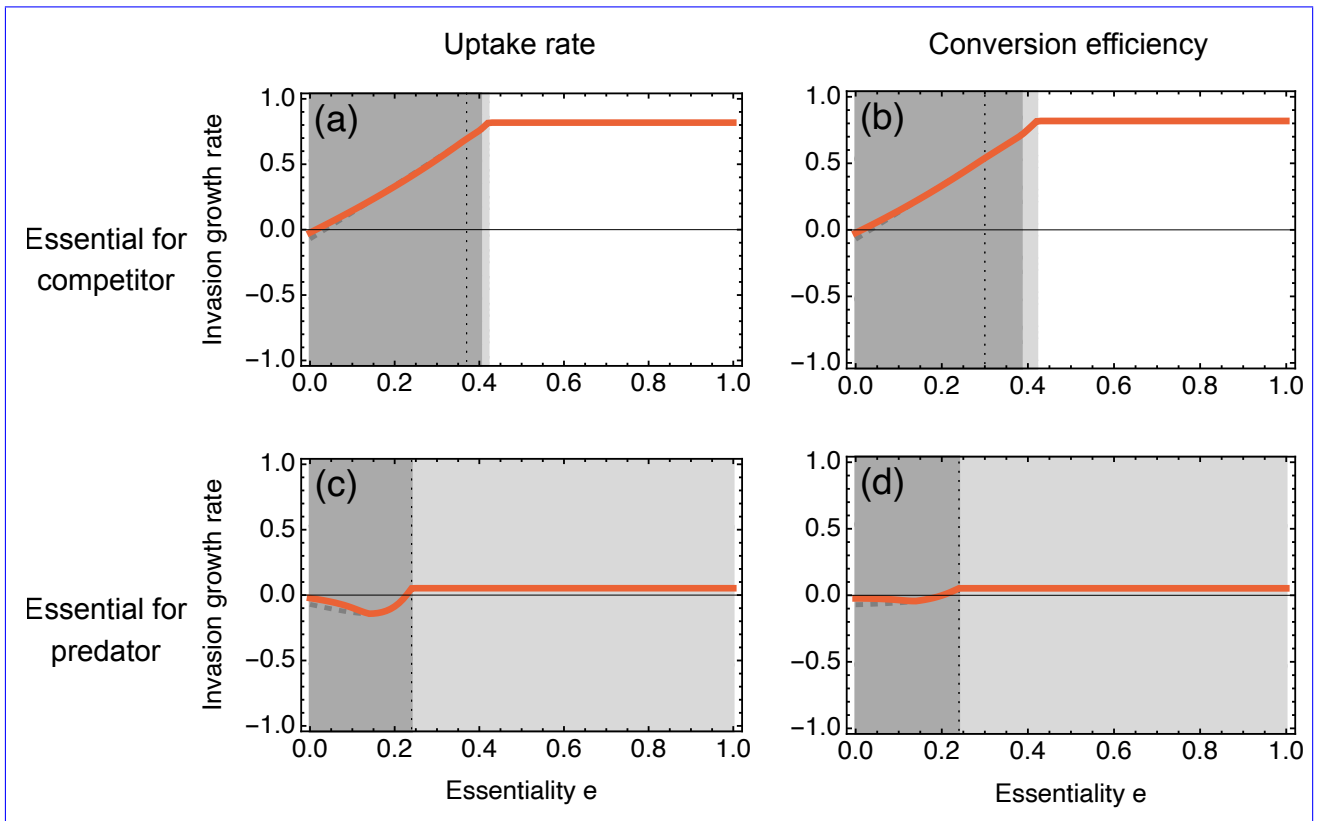


Figure 6 Invasion growth rate for the second food web scenario. Here, the focal prey is more vulnerable to predation but also competitively superior to the competitor. In the absence of the focal prey its essentiality determines the reduction in (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. The plot specifics are identical to Fig. 4.

235 rate at zero essentiality is negative and turns positive for higher essentiality, again resulting in pro-
 236 moted persistence already (Fig. 3a,b). In this scenario, however, higher-order interactions affecting
 237 the predator can result in a negative invasion growth rate, which can become even smaller if the
 238 essentiality becomes larger. Here again, an increasing essentiality counteracts the larger vulnerability
 239 of the competitor to predation, and allows the competitor to outcompete the focal prey given its
 240 higher resource competitiveness.

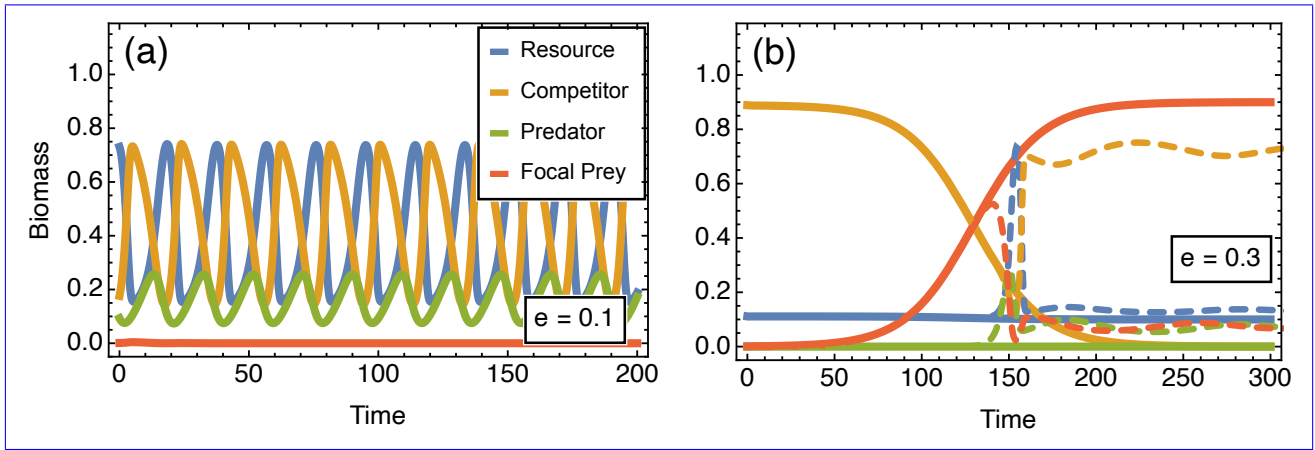


Figure 7 Population dynamics for the second food web scenario when the higher-order interaction targets the resource uptake rate of the shared predator (Fig. 6c). (a) For small essentialities the invasion rate of the focal prey is negative and it thus cannot invade. (b) A larger essentiality ensures the persistence of the focal prey. To obtain these dynamics, I chose the interaction modification according to Eq. 7, integrated the resident community to its stable state and then introduced the focal prey at an initial biomass of $X_{1,0} = 10^{-3}$. In panel (b) the predator goes extinct in the residence community, thus I assumed $Y_0 = 0$ (thick lines). If, however, the predator is reintroduced together with the focal prey ($Y_0 = 10^{-3}$, thin dashed lines), it is supported by the focal prey, re-establishes and mediates the coexistence of both prey types.

Discussion

Higher-order interactions have the potential to shape community structure and dynamics (Grilli et al., 2017; Mayfield & Stouffer, 2017; Terry et al., 2019). In this paper, I showed how the provision of essential resources creates a higher-order interaction that decisively affects the persistence of the focal prey and the resulting community structure. I investigated both the case of essential resource provision to community members from the same trophic level as well as from a higher trophic level. Whether these higher-order interactions in the end ensure persistence depends both on their strength as well as on the food web scenario (see Fig. 2 for a summary of the results). Confirming the expectations, I find in all food web scenarios that a larger essentiality for the competitor can increase the invasion growth rate of the focal prey. In the first and second food web scenario where the invasion growth rate is negative for zero essentiality this leads to a sign-change in the invasion growth rate and thus a promoting effect of essentiality on persistence. In the third

253 and fourth food web scenarios the invasion growth rate of the focal prey is already positive for zero
254 essentiality and only increases further for larger essentiality. Essentiality for the predator can indeed
255 favour the persistence of the focal prey in food web scenarios that permit predator-mediated coexis-
256 tence of the prey species (second food web scenario), but can also be detrimental for persistence if it
257 renders the competitor effectively less vulnerable to predation (fourth food web scenario). Further,
258 I find that essentiality determines the resident community structure, with larger essentiality driving
259 extinct first the predator and then, depending on the higher-order interaction, potentially also the
260 competitor. As seen in the second food web scenario this allows for multiple possible community
261 states, depending on whether the coexistence-mediating predator is re-introduced together with the
262 focal prey. Further, no qualitative differences between higher-order interactions affecting the uptake
263 rate or the conversion efficiency were observed.

264 Experimental support exists for both higher-order interactions that affect the uptake rate or the
265 conversion efficiency. Essential resources affecting the uptake rate could result from adaptive foraging
266 behaviour, as predicted by nutritional geometry ([Raubenheimer & Simpson, 1993](#); [Simpson et al.,](#)
267 [2004](#)), selective feeding ([Buskey, 1997](#); [Elser et al., 2016](#); [Meunier et al., 2016](#); [Eberl et al., 2020](#)),
268 or changed behaviour due to the provision of essential micronutrients, as recently reported for a
269 nematode feeding on larvae of other nematodes ([Akduman et al., 2020](#)). Here, the attack rate
270 of the predatory nematode increased when reared on vitamin B₁₂ producing bacteria compared to
271 B₁₂ deficient controls. However, feeding rate was not increased in this study, so only the prey's
272 loss term would be affected by this higher-order interaction when transferring these results to my
273 model. Another possibility would be generally better physiological conditions that increase fitness,
274 as reported for *Daphnia magna* and vitamin B₁₂ ([Kusari et al., 2017](#)), which could also translate to
275 generally increased activity.

276 The most direct and intuitive mechanism for a higher-order interaction that affects the conversion

277 efficiency of a consumer via essential resource provision is that those lacking essential nutrients that
278 are halting biomass production are directly provided. This is the case in the above example with
279 *Daphnia magna* and vitamin B₁₂ (Keating, 1985), other nutrients like phosphorous (Urabe et al.,
280 2018) or biochemicals (Martin-Creuzburg et al., 2009; Raatz et al., 2017). Similarly, supplementing
281 herbivory with fungivory was found to significantly speed up growth in moth larvae (Eberl et al.,
282 2020). Microbial cross-feeding likely represents the case of higher-order interactions affecting the
283 conversion efficiency of organisms on the same trophic level (D'Souza et al., 2018). In the absence of
284 another carbon source bacteria depend on algal carbon fixation and exudation (Bratbak & Thingstad,
285 1985; Raatz et al., 2018), which was proposed as the mutualistic trade in return for bacterial vitamin
286 B₁₂ provision (Croft et al., 2005) during this type of cross-feeding between different kingdoms.

287 There has been a long history of investigating the effect of higher-order interactions in small ecological
288 interaction networks, such as trait-mediated indirect interactions (Bolker et al., 2003; Werner &
289 Peacor, 2003) or non-trophic interactions (Kéfi et al., 2012), e.g. facilitation (Gross, 2008). The
290 effect of higher-order interactions on community stability is investigated also in larger networks, both
291 empirical (González González et al., 2021) and theoretical, randomly sampled ones (e.g. Arditi et al.,
292 2005; Grilli et al., 2017; Gibbs et al., 2023), and innovative approaches of analyzing their effects have
293 been proposed (Golubski et al., 2016). The effect of trait-mediated indirect interactions and higher-
294 order interactions in general have been shown to depend on many specifics, such as network structure
295 and interaction strengths. In my model, a higher essentiality corresponds to a higher strength of the
296 higher-order interaction. I found that depending on the food web scenario, food-quality-provision-
297 mediated higher-order interactions can be both promoting but also detrimental to persistence and
298 thus community stability, a finding that resonates with this overall complexity. Exploring the effect of
299 multiple, simultaneously occurring higher-order interactions presents an interesting avenue for future
300 research.

301 The provision of essential resources changes the abiotic environment of the competitors or predators
302 via changing the pool of available essential resources. It can be seen as a form of niche construction
303 that is implicitly included via an interaction modification between two biotic food web components
304 (similar to [Kylafis & Loreau, 2011](#); [Oña et al., 2021](#)). Obviously, the niches of predator and com-
305 petitor are impacted directly by the presence of the focal prey. Interestingly, however, this niche
306 construction operates also indirectly in the second food web scenario, as the niche of the focal prey
307 is extended through a feedback loop via predator-mediated coexistence of competitor and focal prey.
308 Bridging theory and experiments on higher-order-interactions is challenging ([Levine et al., 2017](#)).
309 I worked out that essentiality, defined as the reduction of uptake rates or conversion efficiencies
310 when the focal prey is absent, is an appropriate measure to determine the effect of the higher-
311 order interaction on the persistence of the focal prey, particularly when using invasion analysis. One
312 of the benefits from this definition is that the density-dependent functional form of the higher-
313 order interactions does not need to be specified, which largely facilitates experimental approaches
314 of measuring the presence and effect of the higher-order interactions. In my analysis I focussed on
315 the persistence of the focal prey. It should be noted that determining coexistence of species, and
316 not only persistence of a focal species, can be complicated by the existence of multiple stable states
317 (e.g. [Yamamichi et al., 2014](#)) which constrains the interpretation of invasion growth rates ([Grainger
318 et al., 2019](#)).

319 Measuring higher-order interactions experimentally is difficult, however, some advances have been
320 reported that employ different strategies. A first line of research infers the higher-order interactions
321 statistically from community dynamics data (e.g. [Kéfi et al., 2015](#); [Mayfield & Stouffer, 2017](#)).
322 A second, more mechanistic approach aims to disable hypothesized higher-order interactions and
323 compare the outcomes with the non-manipulated scenario. One prominent example is the study by
324 [Wootton \(1993\)](#) where the disguising effect of barnacles for limpets was discovered by removing

325 barnacles partially or completely. Removing the species that initiates the higher-order interaction to
326 quantify the effect of the higher-order interaction however is complicated by other direct and indirect
327 effects that are then also removed, which would lead to false evaluations of the effect size of the
328 higher-order interaction. The essential resource context provides a different way of determining the
329 effect size of higher-order interactions. Experimentally providing the essential resource in excess by
330 supplementation removes its potential to cause higher-order interactions and decreases its essentiality.
331 This approach has been used in investigations of microbial cross-feeding, such as in [Kazamia et al.](#)
332 [\(2012\)](#) and [Hammarlund et al. \(2019\)](#) where supplementation with the essential resource alleviated
333 the dependence on the interaction partner, shifting the coexistence pattern towards the beneficiary
334 of the supplementation. In the context of herbivore limitation by biochemicals, supplementation was
335 used to show the mechanistic basis for the higher-order interaction ([Wacker & Martin-Creuzburg,](#)
336 [2012](#)). In a predator-prey context Bayesian inference from population size time series can be used to
337 obtain uptake rates and conversion efficiencies independently ([Rosenbaum et al., 2019](#)). Applying
338 the inference for different supplementation levels should allow to disentangle whether the essential
339 resource affects the uptake rate or the conversion efficiency of the predator. This may be less feasible
340 for a prey consuming abiotic resources, but even here methods such as isotopic labelling could
341 be used to track uptake and conversion separately. The community-structuring effect of essential
342 resource provision remains to be tested, which requires tracking the population feedback mechanisms
343 over larger time scales of many prey generations, but chemostat or mesocosm experiments will be
344 useful here. The central focus of this article on persistence of the focal prey, however, facilitates
345 experimental validation. As argued before, only the invasion growth rate of the focal prey would have
346 to be obtained for different levels of supplementation with potentially different resident communities.
347 This reduces the time that experimental cultures would have to be operated and avoids experimental
348 difficulties often entailed by long-term observations, ultimately illuminating the potential effect of

349 essential resource provision on prey persistence.

350 **Acknowledgements**

351 The author would like to thank Elias Ehrlich for helpful comments on an earlier draft. Funding for
352 this research was provided by Max Planck Society.

353 **Competing Interests**

354 The author states no competing financial interests.

355 **Data, script and code availability**

356 The Mathematica scripts used for the calculations and for creating the figures have been de-
357 posited both as Mathematica notebooks and also as pdf exports in a zenodo repository with DOI
358 [10.5281/zenodo:8093740.7575588](https://doi.org/10.5281/zenodo.8093740.7575588).

359 **References**

- 360 Abrams, P. A. (1999). Is Predator-Mediated Coexistence Possible in Unstable Systems? *Ecology*,
361 80(2), 608–621. <https://doi.org/10.2307/176639>
- 362 Adamowicz, E. M., Flynn, J., Hunter, R. C., & Harcombe, W. R. (2018). Cross-feeding modulates
363 antibiotic tolerance in bacterial communities. *The ISME Journal*, 12(11), 2723–2735. <https://doi.org/10.1038/s41396-018-0212-z>
- 365 Akduman, N., Lightfoot, J. W., Röseler, W., Witte, H., Lo, W.-S., Rödelsperger, C., & Sommer,
366 R. J. (2020). Bacterial vitamin B12 production enhances nematode predatory behavior. *The ISME*
367 *Journal*, 14(6), 1494–1507. <https://doi.org/10.1038/s41396-020-0626-2>
- 368 Andersen, T., Elser, J. J., & Hessen, D. O. (2004). Stoichiometry and population dynamics: Sto-
369 ichiometry and population dynamics. *Ecology Letters*, 7(9), 884–900. <https://doi.org/10.1111/j.1461-0248.2004.00646.x>
- 370

- 371 Anderson, T. R. & Hessen, D. O. (2005). Threshold elemental ratios for carbon versus phosphorus
372 limitation in *Daphnia*. *Freshwater Biology*, 50(12), 2063–2075. [https://doi.org/10.1111/j.
373 1365-2427.2005.01450.x](https://doi.org/10.1111/j.1365-2427.2005.01450.x)
- 374 Arditi, R., Michalski, J., & Hirzel, A. H. (2005). Rheagogies: Modelling non-trophic effects in food
375 webs. *Ecological Complexity*, 2(3), 249–258. [https://doi.org/10.1016/j.ecocom.2005.04.
376 003](https://doi.org/10.1016/j.ecocom.2005.04.003)
- 377 Azam, F., Fenchel, T., Field, J.G., Gray, J.S., Meyer-Reil, L.A., & Thingstad, F. (1983). The Ecological
378 Role of Water-Column Microbes in the Sea. *Marine Ecology Progress Series*, 10(November 2015),
379 257–263. <https://doi.org/10.3354/meps010257>
- 380 Billick, I. & Case, T. J. (1994). Higher Order Interactions in Ecological Communities: What Are
381 They and How Can They be Detected? *Ecology*, 75(6), 1529–1543. [https://doi.org/10.
382 2307/1939614](https://doi.org/10.2307/1939614)
- 383 Bolker, B., Holyoak, M., Křivan, V., Rowe, L., & Schmitz, O. (2003). Connecting Theoretical and
384 Empirical Studies of Trait-Mediated Interactions. *Ecology*, 84(5), 1101–1114. [https://doi.
385 org/10.1890/0012-9658\(2003\)084\[1101:CTAES0\]2.0.CO;2](https://doi.org/10.1890/0012-9658(2003)084[1101:CTAES0]2.0.CO;2)
- 386 Bratbak, G. & Thingstad, T. F. (1985). Phytoplankton-bacteria interactions: An apparent paradox?
387 Analysis of a model system with both competition and commensalism. *Marine Ecology Progress
388 Series*, 25, 23–30. [https://doi.org/10.1016/0198-0254\(86\)91170-2](https://doi.org/10.1016/0198-0254(86)91170-2)
- 389 Burian, A., Nielsen, J. M., & Winder, M. (2020). Food quantity–quality interactions and their impact
390 on consumer behavior and trophic transfer. *Ecological Monographs*, 90(1). [https://doi.org/
391 10.1002/ecm.1395](https://doi.org/10.1002/ecm.1395)
- 392 Buskey, E. J. (1997). Behavioral components of feeding selectivity of the heterotrophic dinoflagellate
393 *Protoperidinium pellucidum*. *Marine Ecology Progress Series*, 153(1-3), 77–89. [https://doi.
394 org/10.3354/meps153077](https://doi.org/10.3354/meps153077)
- 395 Chesson, P. (1994). Multispecies Competition in Variable Environments. *Theoretical Population
396 Biology*, 45(3), 227–276. <https://doi.org/10.1006/tpbi.1994.1013>
- 397 Croft, M. T., Lawrence, A. D., Raux-Deery, E., Warren, M. J., & Smith, A. G. (2005). Algae
398 acquire vitamin B12 through a symbiotic relationship with bacteria. *Nature*, 438(7064), 90–93.
399 <https://doi.org/10.1038/nature04056>
- 400 Douglas, A. E. (2015). Multiorganismal Insects: Diversity and Function of Resident Mi-
401 croorganisms. *Annual Review of Entomology*, 60(1), 17–34. [https://doi.org/10.1146/
402 annurev-ento-010814-020822](https://doi.org/10.1146/annurev-ento-010814-020822)
- 403 D'Souza, G., Shitut, S., Preussger, D., Yousif, G., Waschina, S., & Kost, C. (2018). Ecology and
404 evolution of metabolic cross-feeding interactions in bacteria. *Natural Product Reports*, 35(5),
405 455–488. <https://doi.org/10.1039/c8np00009c>
- 406 Eberl, F., Fernandez de Bobadilla, M., Reichelt, M., Hammerbacher, A., Gershenzon, J., & Unsicker,
407 S. B. (2020). Herbivory meets fungivory: Insect herbivores feed on plant pathogenic fungi for their
408 own benefit. *Ecology Letters*, 23(7), 1073–1084. <https://doi.org/10.1111/ele.13506>

- 409 Ehrlich, E., Becks, L., & Gaedke, U. (2017). Trait-fitness relationships determine how trade-off
410 shapes affect species coexistence. *Ecology*, 98(12), 3188–3198. [ecy.2047](https://doi.org/10.1002/
411 <a href=)
- 412 Ehrlich, E. & Gaedke, U. (2018). Not attackable or not crackable – How pre- and post-attack defenses
413 with different competition costs affect prey coexistence and population dynamics. *Ecology and*
414 *Evolution*, 8(13), 6625–6637. <https://doi.org/10.1002/ece3.4145>
- 415 Ellner, S. P., Snyder, R. E., Adler, P. B., & Hooker, G. (2019). An expanded modern coexistence
416 theory for empirical applications. *Ecology Letters*, 22(1), 3–18. [ele.13159](https://doi.org/10.1111/
417 <a href=)
- 418 Elser, J., Kyle, M., Learned, J., McCrackin, M., Peace, A., & Steger, L. (2016). Life on the stoichio-
419 metric knife-edge: Effects of high and low food C:P ratio on growth, feeding, and respiration in
420 three *Daphnia* species. *Inland Waters*, 6(2), 136–146. <https://doi.org/10.5268/IW-6.2.908>
- 421 Elser, J. J., Dobberfuhl, D. R., MacKay, N. A., & Schampel, J. H. (1996). Organism Size, Life
422 History, and N:P Stoichiometry. *BioScience*, 46(9), 674–684. [1312897](https://doi.org/10.2307/
423 <a href=)
- 424 Filipiak, M., Kuszewska, K., Asselman, M., Denisow, B., Stawiarz, E., Woyciechowski, M., & Weiner,
425 J. (2017). Ecological stoichiometry of the honeybee: Pollen diversity and adequate species compo-
426 sition are needed to mitigate limitations imposed on the growth and development of bees by pollen
427 quality. *PLOS ONE*, 12(8), e0183236. <https://doi.org/10.1371/journal.pone.0183236>
- 428 Gaedke, U., Hochstädter, S., & Straile, D. (2002). Interplay between energy limitation and nutritional
429 deficiency: Empirical data and food web models. *Ecological Monographs*, 72(2), 251–270. [https://doi.org/10.1890/0012-9615\(2002\)072\[0251:IBELAN\]2.0.CO;2](https://doi.org/10.1890/0012-9615(2002)072[0251:IBELAN]2.0.CO;2)
- 431 Gibbs, T., Gellner, G., Levin, S. A., McCann, K. S., Hastings, A., & Levine, J. M. (2023). Can
432 higher-order interactions resolve the species coexistence paradox? *bioRxiv*. <https://doi.org/10.1101/2023.06.19.545649>
- 434 Golubski, A. J., Westlund, E. E., Vandermeer, J., & Pascual, M. (2016). Ecological Networks over
435 the Edge: Hypergraph Trait-Mediated Indirect Interaction (TMII) Structure. *Trends in Ecology*
436 *& Evolution*, 31(5), 344–354. <https://doi.org/10.1016/j.tree.2016.02.006>
- 437 González González, C., Mora Van Cauwelaert, E., Boyer, D., Perfecto, I., Vandermeer, J., & Benítez,
438 M. (2021). High-order interactions maintain or enhance structural robustness of a coffee agroec-
439 osystem network. *Ecological Complexity*, 47, 100951. <https://doi.org/10.1016/j.ecocom.2021.100951>
- 441 Gore, J., Youk, H., & Van Oudenaarden, A. (2009). Snowdrift game dynamics and facultative
442 cheating in yeast. *Nature*, 459(7244), 253–256. <https://doi.org/10.1038/nature07921>
- 443 Grainger, T. N., Levine, J. M., & Gilbert, B. (2019). The Invasion Criterion: A Common Currency
444 for Ecological Research. *Trends in Ecology & Evolution*, 34(10), 925–935. <https://doi.org/10.1016/j.tree.2019.05.007>
- 446 Grilli, J., Barabás, G., Michalska-Smith, M. J., & Allesina, S. (2017). Higher-order interactions
447 stabilize dynamics in competitive network models. *Nature*, 548(7666), 210–213. <https://doi.org/10.1038/nature23273>
- 448

- 449 Gross, K. (2008). Positive interactions among competitors can produce species-rich communities.
450 *Ecology Letters*, 11(9), 929–936. <https://doi.org/10.1111/j.1461-0248.2008.01204.x>
- 451 Guo, F., Kainz, M. J., Sheldon, F., & Bunn, S. E. (2016). The importance of high-quality algal food
452 sources in stream food webs - current status and future perspectives. *Freshwater Biology*, 61(6),
453 815–831. <https://doi.org/10.1111/fwb.12755>
- 454 Hammarlund, S. P., Chacón, J. M., & Harcombe, W. R. (2019). A shared limiting resource leads
455 to competitive exclusion in a cross-feeding system. *Environmental Microbiology*, 21(2), 759–771.
456 <https://doi.org/10.1111/1462-2920.14493>
- 457 Herren, C. M. (2020). Disruption of cross-feeding interactions by invading taxa can cause invasional
458 meltdown in microbial communities. *Proceedings of the Royal Society B: Biological Sciences*,
459 287(1927), 20192945. <https://doi.org/10.1098/rspb.2019.2945>
- 460 Iwabuchi, T. & Urabe, J. (2012). Food quality and food threshold: Implications of food stoichiometry
461 to competitive ability of herbivore plankton. *Ecosphere*, 3(6), art51–art51. <https://doi.org/10.1890/ES12-00098.1>
- 463 Jones, L. E. & Ellner, S. P. (2007). Effects of rapid prey evolution on predator–prey cycles. *Journal*
464 *of Mathematical Biology*, 55(4), 541–573. <https://doi.org/10.1007/s00285-007-0094-6>
- 465 Kazamia, E., Czesnick, H., Nguyen, T. T. V., Croft, M. T., Sherwood, E., Sasso, S., Hodson, S. J.,
466 Warren, M. J., & Smith, A. G. (2012). Mutualistic interactions between vitamin B12-dependent
467 algae and heterotrophic bacteria exhibit regulation. *Environmental Microbiology*, 14(6), 1466–
468 1476. <https://doi.org/10.1111/j.1462-2920.2012.02733.x>
- 469 Keating, K. I. (1985). The Influence of Vitamin B12 Deficiency on the Reproduction of *Daphnia*
470 *Pulex* Leydig (Cladocera). *Journal of Crustacean Biology*, 5(1), 130–136. <https://doi.org/10.2307/1548225>
- 472 Kéfi, S., Berlow, E. L., Wieters, E. A., Joppa, L. N., Wood, S. A., Brose, U., & Navarrete, S. A.
473 (2015). Network structure beyond food webs: Mapping non-trophic and trophic interactions on
474 Chilean rocky shores. *Ecology*, 96(1), 291–303. <https://doi.org/10.1890/13-1424.1>
- 475 Kéfi, S., Berlow, E. L., Wieters, E. A., Navarrete, S. A., Petchey, O. L., Wood, S. A., Boit, A., Joppa,
476 L. N., Lafferty, K. D., Williams, R. J., Martinez, N. D., Menge, B. A., Blanchette, C. A., Iles,
477 A. C., & Brose, U. (2012). More than a meal... integrating non-feeding interactions into food webs.
478 *Ecology Letters*, 15(4), 291–300. <https://doi.org/10.1111/j.1461-0248.2011.01732.x>
- 479 Koedooder, C., Stock, W., Willems, A., Mangelinckx, S., De Troch, M., Vyverman, W., & Sabbe, K.
480 (2019). Diatom-Bacteria Interactions Modulate the Composition and Productivity of Benthic Di-
481 atom Biofilms. *Frontiers in Microbiology*, 10. <https://doi.org/10.3389/fmicb.2019.01255>
- 482 Koussoroplis, A. M., Schällicke, S., Raatz, M., Bach, M., & Wacker, A. (2019). Feeding in the
483 frequency domain: Coarser-grained environments increase consumer sensitivity to resource vari-
484 ability, covariance and phase. *Ecology Letters*, 22(7), 1104–1114. <https://doi.org/10.1111/ele.13267>
- 486 Kusari, F., O'Doherty, A. M., Hodges, N. J., & Wojewodzcic, M. W. (2017). Bi-directional effects
487 of vitamin B 12 and methotrexate on *Daphnia magna* fitness and genomic methylation. *Scientific*
488 *Reports*, 7(1), 11872. <https://doi.org/10.1038/s41598-017-12148-2>

- 489 Kylafis, G. & Loreau, M. (2008). Ecological and evolutionary consequences of niche construction
490 for its agent. *Ecology Letters*, 11(10), 1072–1081. [https://doi.org/10.1111/j.1461-0248.](https://doi.org/10.1111/j.1461-0248.2008.01220.x)
491 [2008.01220.x](https://doi.org/10.1111/j.1461-0248.2008.01220.x)
- 492 Kylafis, G. & Loreau, M. (2011). Niche construction in the light of niche theory. *Ecology Letters*,
493 14(2), 82–90. <https://doi.org/10.1111/j.1461-0248.2010.01551.x>
- 494 Laland, K., Matthews, B., & Feldman, M. W. (2016). An introduction to niche construction theory.
495 *Evolutionary Ecology*, 30(2), 191–202. <https://doi.org/10.1007/s10682-016-9821-z>
- 496 Levine, J. M., Bascompte, J., Adler, P. B., & Allesina, S. (2017). Beyond pairwise mechanisms of
497 species coexistence in complex communities. *Nature*, 546(7656), 56–64. [https://doi.org/10.](https://doi.org/10.1038/nature22898)
498 [1038/nature22898](https://doi.org/10.1038/nature22898)
- 499 MacArthur, R. & Levins, R. (1967). The Limiting Similarity, Convergence, and Divergence of
500 Coexisting Species. *The American Naturalist*, 101(921), 377–385. [https://doi.org/10.1086/](https://doi.org/10.1086/282505)
501 [282505](https://doi.org/10.1086/282505)
- 502 Martin-Creuzburg, D., Sperfeld, E., & Wacker, A. (2009). Colimitation of a freshwater herbivore by
503 sterols and polyunsaturated fatty acids. *Proceedings of the Royal Society B: Biological Sciences*,
504 276(1663), 1805–1814. <https://doi.org/10.1098/rspb.2008.1540>
- 505 Mayfield, M. M. & Stouffer, D. B. (2017). Higher-order interactions capture unexplained complexity
506 in diverse communities. *Nature Ecology and Evolution*, 1(3), 1–7. [https://doi.org/10.1038/](https://doi.org/10.1038/s41559-016-0062)
507 [s41559-016-0062](https://doi.org/10.1038/s41559-016-0062)
- 508 Meunier, C. L., Boersma, M., Wiltshire, K. H., & Malzahn, A. M. (2016). Zooplankton eat what
509 they need: Copepod selective feeding and potential consequences for marine systems. *Oikos*,
510 125(1), 50–58. <https://doi.org/10.1111/oik.02072>
- 511 Miele, V., Guill, C., Ramos-Jiliberto, R., & Kéfi, S. (2019). Non-trophic interactions strengthen the
512 diversity—functioning relationship in an ecological bioenergetic network model. *PLoS Computa-*
513 *tional Biology*, 15(8). <https://doi.org/10.1371/journal.pcbi.1007269>
- 514 Muller, E. B., Nisbet, R. M., Kooijman, S. A. L. M., Elser, J. J., & McCauley, E. (2001). Sto-
515 ichiometric food quality and herbivore dynamics. *Ecology Letters*, 4(6), 519–529. <https://doi.org/10.1046/j.1461-0248.2001.00240.x>
- 517 Oña, L., Giri, S., Avermann, N., Kreienbaum, M., Thormann, K. M., & Kost, C. (2021). Obligate
518 cross-feeding expands the metabolic niche of bacteria. *Nature Ecology & Evolution*, 5(9), 1224–
519 1232. <https://doi.org/10.1038/s41559-021-01505-0>
- 520 Oña, L. & Kost, C. (2022). Cooperation increases robustness to ecological disturbance in microbial
521 cross-feeding networks. *Ecology Letters*, 25(6), 1410–1420. [https://doi.org/10.1111/ele.](https://doi.org/10.1111/ele.14006)
522 [14006](https://doi.org/10.1111/ele.14006)
- 523 Pomeroy, L., leB. Williams, P. J., Azam, F., & Hobbie, J. (2007). The Microbial Loop. *Oceanography*,
524 20(2), 28–33. <https://doi.org/10.5670/oceanog.2007.45>
- 525 Raatz, M., Gaedke, U., & Wacker, A. (2017). High food quality of prey lowers its risk of extinction.
526 *Oikos*, 126(10), 1501–1510. <https://doi.org/10.1111/oik.03863>

- 527 Raatz, M., Schällicke, S., Sieber, M., Wacker, A., & Gaedke, U. (2018). One man's trash is another
528 man's treasure—the effect of bacteria on phytoplankton-zooplankton interactions in chemostat
529 systems: Bacteria in chemostat experiments. *Limnology and Oceanography: Methods*, 16(10),
530 629–639. <https://doi.org/10.1002/lom3.10269>
- 531 Raatz, M., van Velzen, E., & Gaedke, U. (2019). Co-adaptation impacts the robustness of preda-
532 tor–prey dynamics against perturbations. *Ecology and Evolution*, 9(7), 3823–3836. <https://doi.org/10.1002/ece3.5006>
- 533
- 534 Raubenheimer, D. & Simpson, S. J. (1993). The geometry of compensatory feeding in the locust.
535 *Animal Behaviour*, 45(5), 953–964. <https://doi.org/10.1006/anbe.1993.1114>
- 536 Rosenbaum, B., Raatz, M., Weithoff, G., Fussmann, G. F., & Gaedke, U. (2019). Estimating
537 Parameters From Multiple Time Series of Population Dynamics Using Bayesian Inference. *Frontiers*
538 *in Ecology and Evolution*, 6, 234–234. <https://doi.org/10.3389/fevo.2018.00234>
- 539 Schade, J. D., Kyle, M., Hobbie, S. E., Fagan, W. F., & Elser, J. J. (2003). Stoichiometric
540 tracking of soil nutrients by a desert insect herbivore. *Ecology Letters*, 6(2), 96–101. <https://doi.org/10.1046/j.1461-0248.2003.00409.x>
- 541
- 542 Schällicke, S., Sobisch, L.-Y., Martin-Creuzburg, D., & Wacker, A. (2019). Food quantity–quality co-
543 limitation: Interactive effects of dietary carbon and essential lipid supply on population growth of a
544 freshwater rotifer. *Freshwater Biology*, 64(5), 903–912. <https://doi.org/10.1111/fwb.13272>
- 545 Simpson, S. J., Sibly, R. M., Lee, K. P., Behmer, S. T., & Raubenheimer, D. (2004). Optimal
546 foraging when regulating intake of multiple nutrients. *Animal Behaviour*, 68(6), 1299–1311.
547 <https://doi.org/10.1016/j.anbehav.2004.03.003>
- 548 Singer, M. S., Farkas, T. E., Skorik, C. M., & a Mooney, K. (2012). Tritrophic interactions at
549 a community level: Effects of host plant species quality on bird predation of caterpillars. *The*
550 *American naturalist*, 179(3), 363–374. <https://doi.org/10.1086/664080>
- 551 Soria-Dengg, S., Reissbrodt, R., & Horstmann, U. (2001). Siderophores in marine coastal waters
552 and their relevance for iron uptake by phytoplankton: Experiments with the diatom *Phaeodacty-*
553 *lum tricornutum*. *Marine Ecology Progress Series*, 220, 73–82. [https://doi.org/10.3354/](https://doi.org/10.3354/meps220073)
554 [meps220073](https://doi.org/10.3354/meps220073)
- 555 Stiefs, D., van Voorn, G. A. K., Kooi, B. W., Feudel, U., & Gross, T. (2010). Food Quality in
556 Producer-Grazer Models: A Generalized Analysis. *The American Naturalist*, 176(3), 367–380.
557 <https://doi.org/10.1086/655429>
- 558 Suleiman, M., Zecher, K., Yücel, O., Jagmann, N., & Philipp, B. (2016). Interkingdom Cross-
559 Feeding of Ammonium from Marine Methylamine-Degrading Bacteria to the Diatom *Phaeo-*
560 *dactylum tricornutum*. *Applied and Environmental Microbiology*, 82(24), 7113–7122. <https://doi.org/10.1128/AEM.01642-16>
- 561
- 562 Sun, Z., Koffel, T., Stump, S. M., Grimaud, G. M., & Klausmeier, C. A. (2019). Microbial cross-
563 feeding promotes multiple stable states and species coexistence, but also susceptibility to cheaters.
564 *Journal of Theoretical Biology*, 465, 63–77. <https://doi.org/10.1016/j.jtbi.2019.01.009>

- 565 Terry, J. C. D., Morris, R. J., & Bonsall, M. B. (2019). Interaction modifications lead to greater
566 robustness than pairwise non-trophic effects in food webs. *Journal of Animal Ecology*, 88(11),
567 1732–1742. <https://doi.org/10.1111/1365-2656.13057>
- 568 Urabe, J., Shimizu, Y., & Yamaguchi, T. (2018). Understanding the stoichiometric limitation of
569 herbivore growth: The importance of feeding and assimilation flexibilities. *Ecology Letters*, 21(2),
570 197–206. <https://doi.org/10.1111/ele.12882>
- 571 Wacker, A. & Martin-Creuzburg, D. (2012). Biochemical nutrient requirements of the rotifer Bra-
572 chionus calyciflorus: Co-limitation by sterols and amino acids. *Functional Ecology*, 26(5), 1135–
573 1143. <https://doi.org/10.1111/j.1365-2435.2012.02047.x>
- 574 Werner, E. E. & Peacor, S. D. (2003). A review of trait-mediated indirect interactions in ecological
575 communities. *Ecology*, 84(5), 1083–1100.
- 576 Wolfram Research, Inc. (2019). *Mathematica, Version 12.0*. Wolfram Research, Inc., Champaign,
577 IL.
- 578 Wootton, J. T. (1993). Indirect Effects and Habitat Use in an Intertidal Community: Interaction
579 Chains and Interaction Modifications. *The American Naturalist*, 141(1), 71–89. <https://doi.org/10.1086/285461>
- 580
- 581 Wootton, J. T. (2002). Indirect effects in complex ecosystems: Recent progress and future chal-
582 lenges. *Journal of Sea Research*, 48(2), 157–172. [https://doi.org/10.1016/S1385-1101\(02\)](https://doi.org/10.1016/S1385-1101(02)00149-1)
583 [00149-1](https://doi.org/10.1016/S1385-1101(02)00149-1)
- 584 Yamamichi, M., Yoshida, T., & Sasaki, A. (2014). Timing and propagule size of invasion determine
585 its success by a time-varying threshold of demographic regime shift. *Ecology*, 95(8), 2303–2315.
586 <https://doi.org/10.1890/13-1527.1>

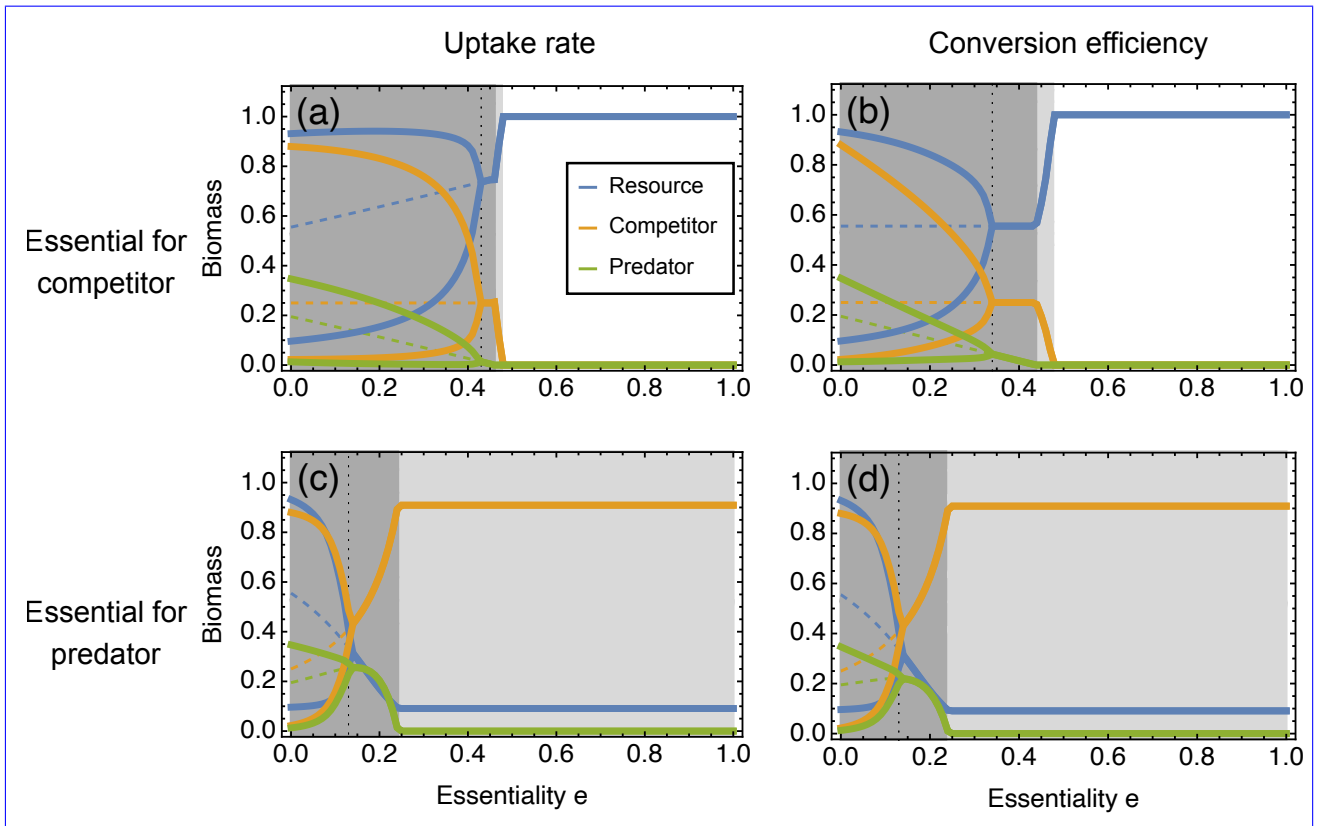


Figure A1 Resident community states for the first food web scenario. In the absence of the focal prey its essentiality determines the reduction in (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator, which affects their community dynamics. Full lines represent the minima and maxima of one population cycle, if the population is cycling, or otherwise the steady state biomass. The vertical dotted line indicates the bifurcation point. Population dynamics were defined as cyclic if the difference between predator extrema exceeded 10^{-5} . During cycles, the unstable fixed point is indicated by the dashed line. As in Fig. 4, the grey shading indicates the states of the resident community. For dark shading both predator and competitor coexist, for light-grey shading only the competitor persists and for no shading only the resource remains.

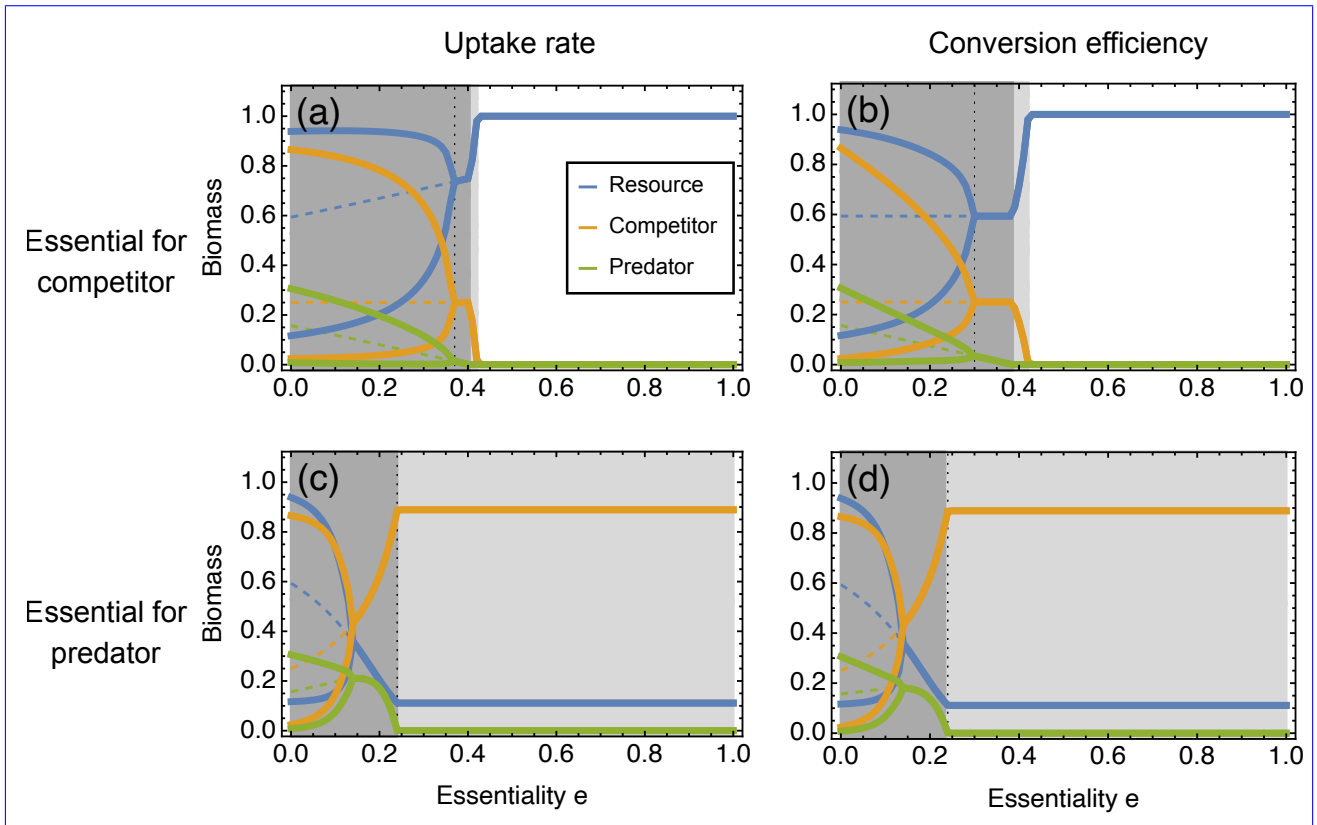


Figure A2 Resident community states for the second food web scenario. Here, the focal prey is more vulnerable to predation but also the superior competitor for the abiotic resource. In the absence of the focal prey its essentiality determines the reduction in (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. The plot specifics are identical to Fig. A1.

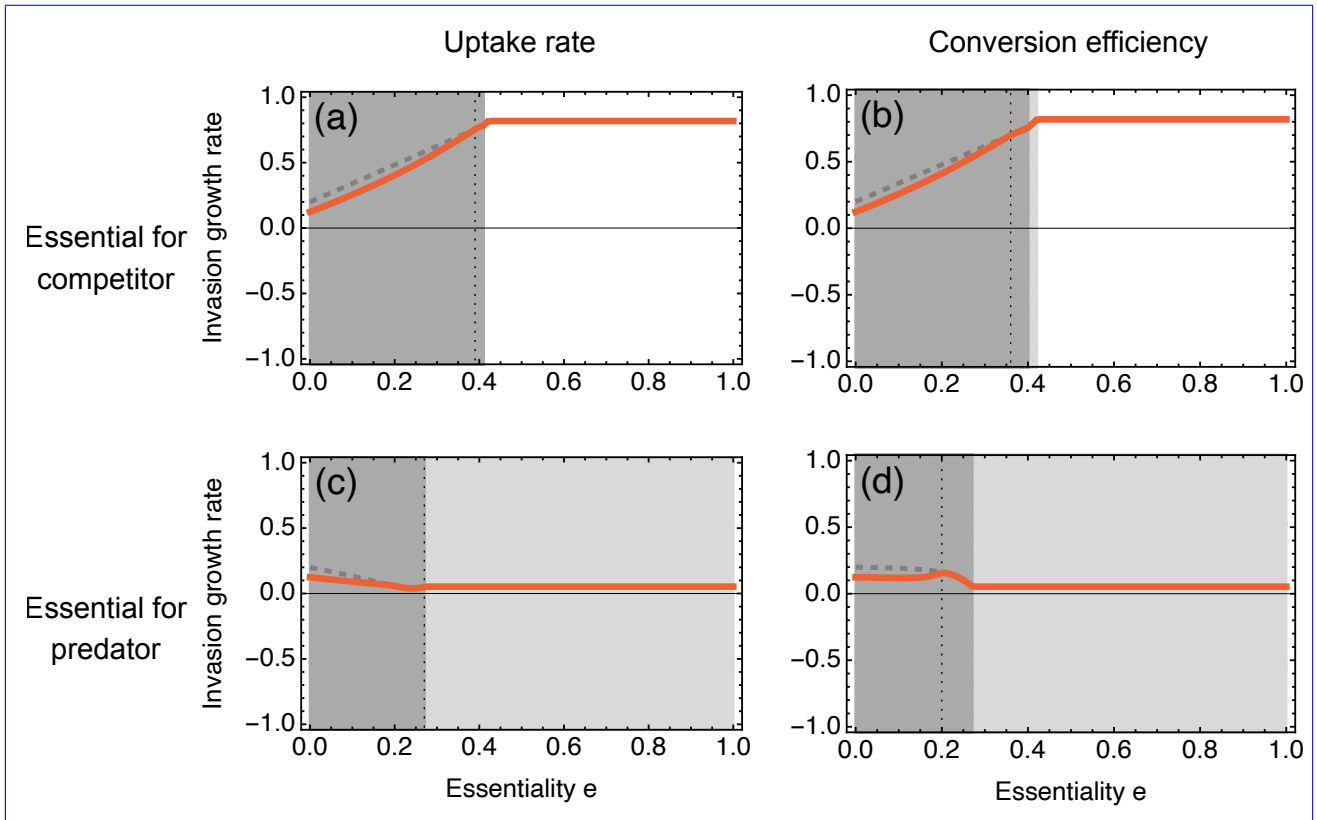


Figure A3 Invasion growth rate of the focal prey for the third food web scenario (Fig. 2c). Here, the focal prey is less vulnerable to predation and more competitive for the resource than the competitor. Essential resource provisioning affects (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. Further plot specifics are identical to Figs. 4 and 6.

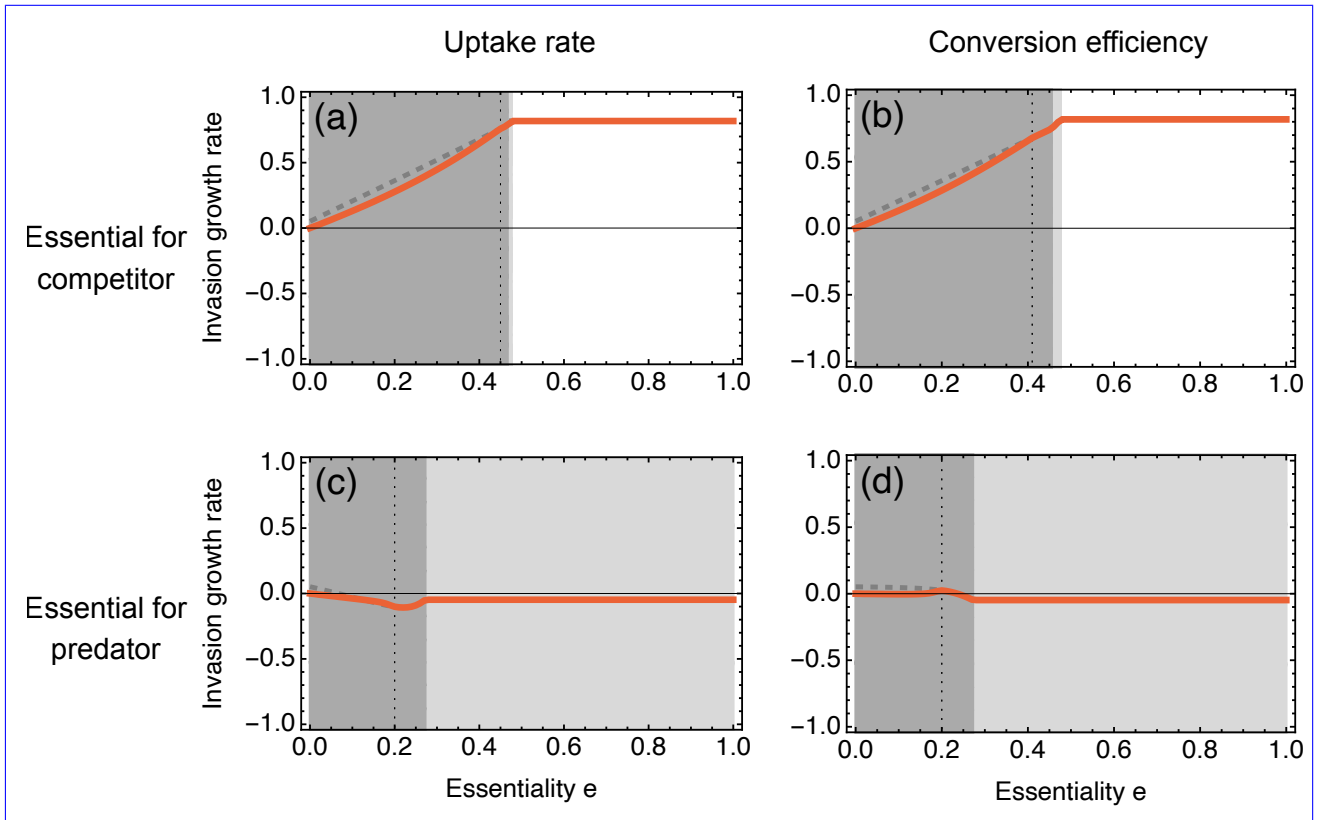


Figure A4 Invasion growth rate of the focal prey for the fourth food web scenario (Fig. 2d). Here, the focal prey is less vulnerable to predation and less competitive for the resource than the competitor. Essential resource provisioning affects (a) the uptake rate or (b) the conversion efficiency of the competitor, or (c) the uptake rate or (d) the conversion efficiency of the shared predator. Further plot specifics are identical to Figs. 4 and 6.