Niche complementarity between pollinators increases

2 <u>community-level</u> plant reproductive success

- 3 Ainhoa Magrach¹, Francisco P. Molina², Ignasi Bartomeus²
- 5 Short running title: Network structure effects on ecosystem functioning
- 6 ¹Basque Centre for Climate Change-BC3, Edif. Sede 1, 1°, Parque Científico UPV-EHU,
- 7 Barrio Sarriena s/n, 48940, Leioa, Spain
- 8 ²Estación Biológica de Doñana (EBD-CSIC), Avda. Américo Vespucio 26, Isla de la
- 9 Cartuja, 41092, Sevilla, Spain
- 10 Type of article: Letter

- Words in Abstract: 170 words
- Words in main text (excluding abstract, acknowledgements, references, table and figure
- 13 legends): <u>5,302</u>
- 14 Number of references: 60
- Number of figures, tables and text boxes: 7
- 16 Corresponding author: ainhoa.magrach@bc3research.org, nacho.bartomeus@gmail.com
- 17 **Keywords**: nestedness, niche complementary, fruit set, pollination, plant-pollinator
- 18 interactions

Abstract

Declines in pollinator diversity and abundance have been reported across different
regions, with implications for the reproductive success of plant species. However,
research has focused primarily on pairwise plant-pollinator interactions, largely
overlooking community-level dynamics. Yet species do not interact in isolation, they are
embedded within larger networks whose structure can affect pollinator functional roles
and, ultimately, the pollination services they deliver to plants. Here, we present one of the
first efforts linking pollinator visitation to plant reproduction from a community-wide
perspective using a well-replicated dataset encompassing 16 well-resolved plant-
pollinator networks and data on reproductive success for 19 plant species from
Mediterranean shrub ecosystems. We find that <u>models including</u> information on simple
visitation metrics alone are good in explaining the variability in reproductive success
observed. However, insights into the mechanisms through which differences in pollinator
diversity translate into changes in reproductive success require additional information on
<u>network</u> structure, particularly that reflecting niche complementarity between pollinators.
Specifically, we find a positive effect of increasing niche complementarity between
pollinators on plant reproductive success.

Introduction

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

Pollinators provide key services to plants by facilitating pollen flow between individuals. The recent declining trends found for some pollinator species in some regions of the planet (Potts et al. 2010, Bartomeus et al. 2019) have led many researchers to focus on the functional impacts of these changes in pollinator diversity, with a major focus being placed on the consequences for plant reproductive success (Biesmeijer et al. 2006). Many research efforts have targeted the reproductive success of individual plant species (Albrecht et al. 2012; Thomson 2019), and used relatively simple visitation metrics (e.g., the number of pollinator species that visit a plant or the number of visits they perform) to explain the differences observed across different plant individuals. Contrastingly, community-level analyses remain scarce (Bennett et al. 2018). Yet plants and pollinators do not interact in isolation, but rather are embedded within larger networks of interactions encompassing other plant and pollinator species. We are thus missing an important part of the picture, which includes the direct interactions between the whole ensemble of plants and pollinators, but also the indirect interactions between species within one guild (e.g., plants) through their shared resources (Pauw 2013; Carvalheiro et al. 2014; Lázaro et al. 2014; Mayfield & Stouffer 2017; Johnson & Bronstein 2019). Understanding how changes in pollinator diversity and interaction structure affect whole community assemblages is thus a major challenge that requires attention. The few pollination studies that have analysed the effects of changing pollinator diversity for reproductive success at the community level have done so using mainly experimental setups. As an example, a study that experimentally recreated a plant

community with 9 plant species and differing levels of pollinator diversity across different enclosures, found that not only pollinator species diversity had an effect for average reproductive success, but that plant-pollinator interaction structure also had an important effect (Fründ et al. 2013). In particular, these authors found that niche complementarity between pollinators, in terms of plant species and temperature coverage (a measure of the overlap in the use of plant resources and optimum temperature activity) had a positive effect for average seed set at the community level (Fründ et al. 2013). This provides added information on the effects of biodiversity for ecosystem functioning, suggesting that not only the diversity of species present, but also the diversity of roles and ways in which a community is structured, are determinant factors. Indeed, theoretical research has long suggested that the structure of multitrophic communities has an effect for ecosystem functioning (reviewed in Thompson et al. 2012). This line of research, rooted in niche theory and revamped by food-web studies (MacArthur & Levins 1967; May & MacArthur 1972, Tilman 1982, Godoy et al 2018), has greatly advanced theory, but these ideas have not yet been tested using empirical data (but see Poisot et al. 2013). Specifically, a major knowledge gap resides in understanding which aspects of structure determine which aspects of function (Thompson et al. 2012). This is because although a network perspective has promised to encapsulate complex ecological mechanisms occurring at the community level – such as indirect interactions (Holt 1977, Abrams et al 1998) or niche overlap (Woodward & Hildrew 2002)- less attention has been given to the ways in which these mechanisms relate to observed ecosystem processes (Blüthgen 2010). In contrast, we are now at a point in which there is considerable understanding on the attributes characterizing mutualistic interaction

59

60

61

62

63

64

65

66

67

68

69

70

71

72

73

74

75

76

77

78

79

80

networks (Bascompte & Jordano 2007). Especially, in the case of pollination, we have ample knowledge on the attributes that shape these mutualistic interactions at the community level. Amongst them is the prevalence of nested structures, i.e., arrangements where specialist species interact with a subset of the species that generalists interact with (Bascompte et al 2003) and which is thought to promote species diversity (Bastolla et al 2009); or the relatively high extent of complementary specialization at the community scale, which may be directly related to key ecosystem functions (Blüthgen & Klein 2011). However, the mechanisms by which these potential pathways affect plant reproduction remain to be understood (Winfree 2013). The time is thus ripe to use the existing knowledge around plant-pollinator network structures to explore the relationship between network structure and ecosystem functioning empirically, with special emphasis being placed on the underlying ecological mechanisms that drive these relationships.

Here, we present one of the first efforts linking pollinator visitation and plant reproductive success at the community level using empirical data on plant-pollinator interaction networks and plant reproductive success. To this end, we use a well-replicated dataset encompassing 16 well-resolved plant-pollinator interaction networks coupled with data on the reproductive success of 19 plant species recorded in Mediterranean shrub ecosystems. Our study focuses on understanding whether adding information on selected interaction network structure indices to previously used simple visitation metrics (e.g., the number and diversity of pollinator species visiting a plant species) aids in better explaining the differences observed in community-wide reproductive success. In doing so, we conducted our analyses focusing on reproductive success at two different levels:

(i) at the species level by considering the effect of the position of a focal species within

the larger network and its impact on its individual reproductive success, and (ii) at the site
level, by evaluating how attributes that describe the whole site might affect average
values of reproductive success. Specifically, our study focuses on how the interplay
between niche complementarity and redundancy determines reproductive success. Plant
reproductive success requires of the delivery of conspecific pollen and thus of a certain
degree of niche complementarity (Blüthgen & Klein 2011). Yet, greater values of
nestedness, which imply redundancy in species functions, are thought to promote species
diversity (Bastolla et al. 2009) and stability (Thébault & Fontaine 2010) within plant-
pollinator networks. At present, we do not know how either of these network
characteristics affects the functions performed by pollinators. Finally, in addition to
average values, we also evaluate whether <u>network</u> structure helps explain differences in
equity in reproductive success across species within a community, as a measure of
evenness in the pollination service delivered.

Our results suggest that <u>models including information on simple visitation metrics</u> alone are good in explaining the variability observed in reproductive success. However, insights into the mechanisms through which differences in pollinator diversity translate into changes in reproductive success require additional information on network structure, notably information on the complementarity between the functions performed and the niches occupied by different pollinator species. <u>Specifically</u>, we find a positive effect of increasing niche complementarity between pollinators on plant reproductive success.

Methods

Plant pollinator interactions

Our study was conducted in SW Spain within the area of influence of Doñana National
Park, i.e., within the limits of the Natural Space of Doñana as defined by the local
government (Junta de Andalucía, Fig. 1). All sites were located within similar elevations
(ranging from 50 to 150 m a.s.l.), similar habitat and soil types, and presented similar
plant composition (plant mean Sørensen beta-diversity among sites = 0.41), reducing
potential confounding factors. Here, we surveyed 16 Mediterranean woodland patches
with an average distance of 7 km between them (min= 3 km, max= 46.5 km). Each site
was surveyed 7 times during the flowering season of 2015 (from February to May)
following a 100-m x 2 m transect for 30 mins. Along each transect, we identified all plant
species and recorded all the floral visitors that landed on their flowers and touched the
plant's reproductive parts during each 30-min period. Only floral visitors (from now on
referred to as pollinators) that could not be identified in the field were captured, stored
and identified in the laboratory by FPM and experts in the different taxonomic groups
(see acknowledgements). In addition, at each round we conducted 3 minutes of focal
observations recording all floral visitors observed on 3 plant individuals per species
belonging to the 19 most common (based on previous surveys) plant species across the
study area ($\underline{\text{mean} \pm \text{SD:}}$ 6.25 \pm 1.73 species per site). Furthermore, we included some
interactions between plant and pollinator individuals that were not observed during the
sampling but that were opportunistically recorded immediately before or after the
sampling periods, as some of these interactions are difficult to document and might be
important to define network structure (Jordano 2016). These opportunistic interactions
represented 22.96% of all interactions recorded. All surveys were done under similar

149 weather conditions, avoiding windy or rainy days. Surveys were done during mornings 150 and afternoons with the sampling order being established randomly. 151 Plant reproductive success 152 Within each site, we marked between 3 and 12 individuals (mean \pm SD: 6.49 \pm 2.37, 153 Table S2) belonging to 1 to 6 plant species (mean \pm SD: 4.06 \pm 1.69, Table S1), 154 depending on the availability and presence of flowers during the sampling events. For 155 each individual, at the end of the season, we recorded fruit set (i.e. the proportion of 156 flowers that set fruit), the average number of seeds per fruit and the average fruit and 157 seed weight per fruit (1-36 fruits subsampled; mean \pm SD: 11.17 \pm 6.85, Table S3). Our 158 survey included a total of 19 different plant species across our 16 sites. Plants species 159 were selected based on their availability, with sampling being focused on the most 160 abundant plant species. The values at the species level were then averaged per site to 161 calculate unique reproductive success measures at the site level. All plant species depend 162 on pollinators to maximize their reproduction (Table S4). 163 Data analyses 164 In order to evaluate the completeness of our sampling of the pollinator and plant 165 community as well as that of their interactions, we estimated the asymptotic number of 166 species present (Chao et al. 2009), a non-parametric estimator of species richness for 167 abundance data. This estimator includes non-detected species and allowed us to calculate 168 the proportion detected with our original sampling data. We used Chao 1 asymptotic 169 species richness estimators (Chao et al. 2009) and estimated the richness of pollinators, 170 plants and plant–pollinator links accumulated as sampling effort increased up to 100%

sampling coverage using package iNEXT (Hsieh *et al.* 2016) within the R environment (R Development Core Team 2011). We then extracted the values covered by our sampling.

In order to analyse how differences in network structure might affect plant reproductive success, we constructed plant-pollinator interaction networks by pooling the data for the 7 rounds of sampling. We thus obtained one interaction network per site, representing the number of individuals of different pollinator species recorded visiting each different plant species. For each network, we then proceeded to extract a series of relevant network metrics at the species and site levels.

In addition, we checked for potential spatial autocorrelation in our data by means of Mantel correlograms. Autocorrelation values were low for all variables included in our analyses (Figure S1) and hence we treat each site as independent in our analysis.

Species-level network analysis

At the species level, we focused on attributes defining the position of a focal plant species within the larger community. As such, we considered two metrics providing complementary non-redundant information: (i) average niche overlap in terms of pollinators between a focal plant species and <u>each of the</u> other plant species in the community, which <u>estimates</u> the potential indirect interactions between different plant species through shared resources (in this case pollinators), and (ii) centrality, which depicts the importance of the role played by a plant species within the larger community (as resource for a large number of pollinator species) and its contribution to network cohesiveness.

Niche overlap was calculated as the average overlap in pollinator species visiting a focal plant and each of the other plants in the community using the Morisita <u>overlap</u> index, a measure of similarity between two sets of data (Zhang 2016). As a measure of centrality, we used weighted closeness centrality, which represents the number of shortest paths going through a focal plant based on a unipartite projection of the bipartite plant-pollinator network using a weighted representation of the network (Dormann *et al.* 2009). Here, links between plant species represent shared pollinator species.

<u>Site</u>-level network analysis

At the <u>site</u> level, we followed the same logic as the one presented at the species level.

Thus, we also calculated two network metrics providing complementary non-redundant information. In this case we focused on (i) nestedness and (ii) <u>pollinator</u> niche complementarity.

Nestedness is the property by which specialists interact with a subset of the species that generalists interact with (Bascompte *et al.* 2003). Although there is an ongoing debate in the literature, some studies have found that nested networks are more stable and resilient to perturbations because nestedness promotes a greater diversity by minimizing competition among species in a community (Bastolla *et al.* 2009). However, many network attributes vary with network size and complexity (Blüthgen *et al.* 2006). In the case of nestedness, we know it can be affected by network size and connectance (Song *et al.* 2017). An approach that is often used to correct for this, is to use null models and to compare null-model corrected nestedness values across different networks. However, this approach has been recently shown to present the same issues, as z-scores also change with network size and connectance (Song *et al.* 2017). We thus followed the

advice provided by Song et al. (2017) by using a normalized value of the widely-used nestedness metric NODF (Almeida-Neto & Ulrich 2011), $NODF_c$. This normalized value is calculated as $NODF_c = NODF_n/(C * log(S))$, where C is connectance and S is network size. $NODF_n$ is calculated as NODF/max(NODF), which is independent of network size and thus comparable across different networks (Song *et al* 2017). To calculate max(NODF) we used a recently corrected version of the algorithm (Simmons *et al* 2019) in all but three sites, where the condition that the number of links>number of species was not met and thus precluded us from using this new version.

Niche complementarity metrics are important because plant reproductive success depends on the delivery of conspecific pollen and thus of a certain level of specialization or niche divergence (reviewed in Brosi 2016). To calculate niche complementarity, we used a community-level measure defined as the total branch length of a functional dendrogram based on qualitative differences measured using a Euclidean distance in visitor assemblages between plants (Petchey & Gaston 2007; Devoto *et al.* 2012). All network metrics were calculated using package bipartite (Dormann *et al.* 2009).

All of these metrics were calculated using all the data as well as for the subset of the data excluding interactions observed outside of sampling periods. Differences between results are minimal for both and thus we will only present results for the analysis using the full dataset (see Table S12A-H for results removing observations out of transect).

Statistical analyses

In order to evaluate whether adding information on network structure improves our ability to explain differences in reproductive success - both at the species and the site level - we used generalized linear (GLMs) and generalized linear mixed models (GLMMs). In both cases (species and site-level models) we fit two types of models: (i) model 1, that only included simple visitation metrics and (ii) model 2 that additionally included information on network structure. These models are meant to be additive, so that the network metrics included are intended to complement rather than substitute the simple metrics traditionally used.

At the species level, response variables included the fruit set for different individuals of each species analyzed using a binomial distribution, the average number of seeds per fruit analyzed using a normal distribution, and the average values of fruit and seed weight fitted to Poisson distributions. The number of seeds per fruit was centered and scaled (i.e., we subtracted column means and divided by standard deviation) to allow meaningful comparisons across species with contrasting life histories. As explanatory variables, model 1 included pollinator richness, and the total number of visits received by each plant species; while model 2 added the two network attributes calculated at the species level: average plant niche overlap and centrality. For both models, we included plant species nested within site and site as random effects to account for the non-independence of several individuals measured per species and site.

At the <u>site</u> level, we upscaled our species-level analyses. <u>As response variables we</u> had the average reproductive success per site (i.e., average fruit set analyzed using a binomial distribution, average number of seeds per fruit and average fruit and seed weight using a normal distribution). We thus had a single value per site and no random

effects are needed. In this case, model 1 included total pollinator richness and total pollinator abundance (i.e. number of visits received by all plants within the community) as explanatory variables. Model 2, in turn, added information on network structure by including nestedness and pollinator niche complementarity as explanatory variables.

Average values of reproductive success at the <u>site</u> level can be driven to a large extent by a single plant species. Yet, what will determine the persistence of a diverse plant community, is the presence of some sort of "equity" or evenness in reproductive success across the whole community. We therefore calculated a measure of equity in reproductive success at the <u>site</u> level as the proportion of species with normalized (between 0 and 1) average fruit set values that were above <u>the 50th percentile</u>. As any <u>selected threshold is arbitrary, we repeated this using the 25th and 75th percentile thresholds (Byrnes *et al* 2014). We then used the same framework as that used for species and <u>site</u>-level analyses and fit the same models 1 and 2 <u>GLMs but</u> using equity in reproductive success as the response variable <u>and</u> fitting a binomial distribution.</u>

In all cases, we used variance inflation factors to check for collinearity between explanatory variables. Additionally, we ran residual diagnostics to check if model assumptions were met. Then, we used the Akaike Information Criterion (AIC) to compare model performance and complexity. Whenever the difference between the AIC of both models was < 2 ($\triangle AIC < 2$), we considered that both models were equally good (Burnham et al. 2011). All predictor variables were standardized prior to analysis. For every model we also calculate the R^2 value, using the approximation suggested for generalized mixed models when necessary (Nakagawa et al 2017).

Finally, we tested whether the importance of <u>network</u> structure in explaining differences in equity in reproductive success within communities increases with the number of plant species being considered. We expect that when only one plant species is considered, then the importance of <u>network</u> structure will be negligible, while we expect this importance to increase as more plant species are considered (up to a maximum number of 6 species which is the maximum we have measured in our study at a particular site).

282

283

284

285

286

287

288

289

290

291

292

293

294

295

296

297

298

299

300

301

302

303

304

To test this, we ran a simple simulation in which the number of species considered increased at each step and for each step we re-calculated equity in reproductive success. Instead of drawing plant species randomly for each step, we tested all possible combinations for each plant number level and network, as the number of combinations is small (e.g. for n = 3 plant selected out of 6 there is only 20 possible combinations). Then, we tested if the relationship between equity in reproductive success and functional complementarity (given its importance in determining differences in reproductive success, see Results section) changes as a function of the number of plants considered within our simulated communities. To this end, for each level of species number considered, we randomly selected one of the generated equity values across each of the 16 communities and regressed these 16 values against our network level predictor and extracted the model slope estimates. We repeated this process 1,000 times and averaged all slope estimates. We expect that the more plants considered, the larger the resulting average estimates will be. Note that we only interpret the mean effects, as the variance among different plant number of species considered depends on the initial number of possible combinations.

Results

305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

322

323

324

325

Within our sampling we recorded 1,472 plant-pollinator interactions involving 277 pollinator species and 57 plant species. Within the pollinator community the distribution of individuals in different orders was: 87.84% Hymenoptera, 6.78% Diptera, 4.05% Coleoptera and 1.09% Lepidoptera. Our sampling completeness analyses revealed that with our survey we were able to capture 18-62% of pollinator species (average = 35%), 47-98% for plant species (average = 78%) and 13-41% for plant-pollinator links (average = 27%), in line with that obtained with other studies (e.g., Chacoff et al. 2012, Fig. S2). Our values of sampling completeness are slightly smaller in the case of pollinators, probably as a consequence of the great diversity found in the Mediterranean region and within our study area in particular, a hotspot of insect diversity (Nieto et al. 2014). In addition, the fact that we include an extra effort to capture rare interactions observed outside of our main sampling might also increase the number of singletons which directly affect richness estimates. Species-level analyses At the species level, in the case of fruit set, our results show that model 2 shows the best fit to our data (lowest AIC value), and fixed effects explains 4% of the variability observed. In this case, we find a positive effect of a network structure metric, the centrality of a focal plant within the overall network on its fruit set (Table 1, Fig. 2A). For the average number of seeds per fruit, our results show again that model 2

shows the best fit, with fixed effects explaining 4% of the variability observed in our

326 data. In this case, we find a positive effect of the niche overlap between plant species on 327 the number of seeds produced (Table 1B, Fig. 2B). 328 For all other measures of reproductive success considered (i.e., fruit and seed 329 weight), both models had similar fits, with \triangle AIC = 2.2 and 2 respectively. However, 330 none of the variables included within our model explain the differences observed (Tables 331 S5-S6). 332 Site-level analyses 333 At the site level, we find different patterns for fruit set and the number of seeds per fruit 334 as compared to those for fruit and seed weight. In the case of fruit set and the number of 335 seeds per fruit, we find that both model 1 and 2 are equally good in describing the differences observed when penalizing for model complexity (i.e., $\Delta AIC < 2$; Burnham et 336 337 <u>al 2011</u>). This suggests model 2 is a good model despite its added complexity, and 338 actually shows a substantially better predictive ability than model 1 ($R^2 = 0.59$ for model 339 2 versus 0.47 for model 1 in the case of fruit set and $R^2 = 0.52$ for model 2 versus 0.31 340 for model 1 in the case of the number of seeds per fruit) and therefore we will comment 341 results for this model only. In particular, we find that both fruit set and the number of 342 seeds per fruit are positively related to niche complementarity between pollinators (Table 343 2, Fig. 3). Additionally, we find a negative effect of site-level pollinator richness on 344 average fruit set (Table 2A, Fig. 3). 345 Contrastingly, in the case of weight variables (fruit and seed weight), in both 346 cases we find that the best model is model 1, i.e., that only including simple visitation 347 metrics ($R^2 = 0.29$ in the case of fruit weight and 0.51 in the case of seed weight). Here,

we find a consistent positive effect of <u>site</u>-level pollinator <u>richness</u> for both weight descriptors (Tables S<u>7</u>-S<u>8</u>, Fig. <u>4</u>).

Equity in fruitset

When evaluating the effect of differences in community composition and <u>network</u> structure for equity in reproductive success across the different species within a community we find that model 1 is the best model for all the thresholds considered (<u>50th</u>, <u>25th</u> and 75th percentiles). <u>However, none of the variables considered are able to explain differences observed in equity across sites</u> (Tables S<u>9</u>, S<u>10</u>, S<u>11</u>).

Within our simulation evaluating the effect of niche complementarity on equity in reproductive success as more plants within the community are considered, we find that the effect of complementarity becomes more important as the reproductive success of more species is considered (Fig. 5). This importance seems to reach some sort of plateau at 6 species. However, this should be further evaluated, as this is the maximum number of species simultaneously observed in a community for our study, which precludes us from simulating further numbers of species.

Discussion

The existence of relationships between <u>interaction network</u> structure and ecosystem function have been long hypothesized, yet, the specific mechanisms by which structure influences function have remained elusive until now (Thompson *et al.* 2012). Our results show that different aspects of <u>network</u> structure affect different dimensions of <u>ecosystem</u> functioning. In particular, we find that the centrality of a plant species within a community, which measures the number of connections it receives from other species in

the community, has a positive effect for its fruit set. At the <u>site</u> level, we find that <u>greater</u> <u>values of</u> niche complementarity between pollinators <u>result in larger</u> average fruit sets <u>and number of seeds per fruit</u>.

One of the first conclusions we can extract from the fact that in most cases both of the models we considered (i.e., the simple model based on visitation metrics and the more complex one including network structure metrics) were equally good, is that the added complexity of measuring the full network of interactions may not pay off for rapid assessments. Hence, simple visitation metrics, such as pollinator richness, might be enough to describe general patterns (Garibaldi et al. 2013, 2015). Yet, adding network level information may inform us of the potential ecological mechanisms underlying the processes driving the observed patterns.

Consistent with previous experimental (Fontaine *et al.* 2005; Fründ *et al.* 2013), theoretical (Pauw 2013), and empirical studies (Valdovinos *et al.* 2016, Poisot *et al.* 2013), we find that niche complementarity is key in determining differences in reproductive outputs. Indeed, we find that communities where there is less overlap in the niches occupied by pollinator species had greater values of reproductive success, both greater fruit set and larger numbers of seeds per fruit. This therefore reflects the fact that reproductive success in plant species requires the delivery of conspecific pollen and thus of a certain degree of specialization amongst pollinator species on a particular plant resource in order to avoid the negative effects of inter-specific pollen deposition (e.g., pollen loss, Flanagan *et al.* 2009) or interference with conspecific pollen (Morales & Traveset 2008). However, we also find that some level of redundancy in these functions

is needed as revealed by the positive effect of plant niche overlap on the number of seeds per fruit at the species level.

In our study, we did not find an effect of nestedness for reproductive success in any case. This metric, widely used across network analysis, and which is deemed to stabilize <u>natural communities</u> (Bastolla *et al.* 2009 <u>but see</u> James *et al.* 2012), does not seem to play a direct role in ecosystem function measured as plant reproductive success. However, our study is limited to a maximum of six <u>common</u> plant species per community, and including more species, <u>including rare species</u>, might reveal different patterns, in which nestedness and the redundancy it implies might play a more important role.

Site-level plant reproductive success measured as average fruit or seed set across all the species considered, is an important part of the functions delivered by pollinators to plants. However, these average values might be masking a great deal of variability amongst plant species, and thus a nuanced view of the effect of pollinators on whole-plant ensembles is needed. This can be captured by the effect of pollinators on equity in reproductive success across plant species. This aspect ensures that reproductive success is equally distributed amongst a larger number of species, thus contributing to the maintenance of greater species diversity values in natural populations. Indeed, we know that plant species diversity within a community is largely driven by different types of direct and indirect interactions including those amongst plant species (e.g., resource competition, Goldberg & Barton 1992, or facilitation, Bruno et al. 2003), as well as those defining antagonistic (e.g., involving pathogens, Bagchi et al. 2010), or mutualistic interactions (e.g., pollinators, Benadi et al. 2013; Lanuza et al. 2018). However,

equitability in reproductive success across species is seldom taken into account, despite its importance in maintaining genetic diversity and ensuring the resilience of populations to further change.

In the case of equity, we did not find a <u>strong</u> effect of either simple visitation or <u>network</u> structure metrics. However, the results of our simulation on the importance of <u>network</u> structure as the number of plant species considered increases, shows us that this effect increases when more than four plant species are considered. This implies that if we were able to measure reproductive success for all the plant species in all the communities (which is not feasible given constraints in sampling effort), we might find that the effects of <u>network</u> structure on equity might be more prevalent.

One of the unexpected results of our analyses is the strong negative effect of pollinator <u>richness</u> for fruit set at the <u>site</u> level. An explanation to this might be the fact that pollinator <u>richness</u> here includes all the pollinators <u>recorded during our sampling</u> efforts, i.e., it includes species that do not pollinate some of the species whose reproductive success was measured. More complex communities with more pollinators, but also with more plant species (Pearson correlation <u>between plant and pollinator richness = 0.42</u> in our case) may require stabilizing mechanisms that reduce the competition exerted by the dominant plant species. A way to reduce the competition exerted by these dominant species, which are precisely those evaluated in this study, is by reducing their reproductive success (Lanuza et al 2018, Stavert et al 2019). These ideas open the door to exploring the positive or negative effects of the complete pollinator community on full plant species coexistence, which may be determined by density-dependence effects (Benadi & Pauw 2018). In our case, while fruit set is negatively

related to pollinator richness, it is important to note that fruit and seed weight show the opposite relationship, indicating that this density-dependent effect might only be limiting fruit quantity and not fruit quality. Thus, taking into account the densities of co-flowering plant species may be the next step (Vanbergen *et al.* 2014).

Our study illustrates the complexity of linking network structure to ecosystem function empirically, because measuring both structure and function is challenging. For example, there is an ongoing debate as to which network metrics better reflect classic ecological mechanisms, such as niche partitioning or competition (Delmas *et al* 2018). Here, we focus on testing two specific hypotheses, but other structural properties can be explored when more data becomes available. Furthermore, the structure of plant-pollinators networks is dynamic due to ecological and evolutionary reasons, but so far, we are only able to characterize it for single snap-shots. Moreover, different aspects of functioning may be important, such as the presence of non-linear relationships or the need to consider the functioning of both trophic levels (Godoy *et al* 2018). In terms of plant reproductive success and the functions performed by pollinators we can measure different aspects, ranging from pollen deposition (the direct pollinator function), to its final effects on plant fitness. Here, we focus on an intermediate stage including fruit quantity and quality, which is of clear ecological importance.

In summary, our findings show that the analysis of natural communities of interacting species using network analysis not only represents an ideal way of visualizing and grasping the complexity present within these communities. Rather, it also represents a manner of mechanistically understanding differences observed across the reproductive

460 success of individuals and/or species while linking them to potential ecological 461 mechanisms. 462 **Data accessibility** 463 All the data used is available 464 https://zenodo.org/account/settings/github/repository/ibartomeus/BeeFunData and the 465 code used to generate all results can be found at https://doi.org/10.5281/zenodo.3364037. 466 467 Acknowledgements 468 The authors would like to thank Oscar Aguado for identifying pollinator species. We 469 would also like to acknowledge the efforts of the editor Cédric Gaucherel, as well as 470 those of Michael Lattorff, Nicolas Deguines, and three anonymous reviewers who 471 provided very helpful comments and suggestions in a previous version of this manuscript. 472 AM received funding from a Juan de la Cierva (IJCI-2014-22558) and Ikerbasque 473 fellowships. IB acknowledges funding from MSC-PCIG14-GA-2013-631653 BeeFun 474 Project. We thank Doñana's Singular Scientific-Technical Infrastructure (ICTS-RBD) for 475 access to the park.

Table 1. Results of GLMM showing <u>the</u> effect of simple visitation and <u>network</u> structure metrics on A) species-level fruit_set and B) average number of seeds per fruit based on best model selected. Bold letters indicate variables <u>with large effects</u>.

A) Fruit set	Estimate	Std.Error	<u>z.value</u>
(Intercept)	1.72	0.21	8.16
Pollinator richness	<u>-0.01</u>	0.21	<u>-0.07</u>
Total number of visits	<u>0.14</u>	0.25	0.57
Centrality	<u>0.46</u>	<u>0.25</u>	<u>1.81</u>
Plant niche overlap	0.05	0.24	0.20

B) Seeds per fruit	Estimate	StdError	t.value
(Intercept)	0.05	0.14	0.36
Pollinator <u>richness</u>	0.14	0.15	0.90
Total number of visits	-0.01	0.16	-0.04
Centrality	-0.15	0.15	-1.00
Plant niche overlap	0.22	0.17	1.32

Table 2. Results of GLM showing effect of simple visitation and <u>network</u> structure metrics on A) <u>site</u>-level average fruit set and B) <u>site</u>-level average number of seeds per fruit based on best model selected. <u>Bold letters indicate variables with large effects.</u>

A) Fruit set	Estimate	Std. Error	z value
(Intercept)	1.22	0.13	9. <u>1</u> 8
Pollinator richness	-0.75	0.17	-4.35
Total number of visits	-0.1 <u>6</u>	0.16	-1. <u>0</u> 2
Nestedness	0. <u>11</u>	0. <u>16</u>	0. <u>72</u>
Pollinator n iche complementarity	0. <u>29</u>	0.1 <u>8</u>	<u>1.57</u>

B) Seeds per fruit	Estimate	Std. Error	t value
(Intercept)	45.3 <u>7</u>	8. <u>55</u>	5. <u>31</u>
Pollinator richness	<u>3</u> . <u>26</u>	1 <u>2</u> . <u>18</u>	0. <u>27</u>
Total number of visits	8. <u>38</u>	<u>9.99</u>	0.8 <u>4</u>
Nestedness	<u>-10</u> . <u>94</u>	1 <u>0.21</u>	<u>-1</u> . <u>07</u>
Pollinator n iche complementarity	29.51	13.31	2.21

491 Figure legends 492 Figure 1. Map showing location of 16 study sites. Inset shows location of study area 493 within SW Spain. 494 Figure 2. Partial residual plot showing the effect of a single predictor for the relationship 495 between A) plant species centrality and fruit set for each of the plant species considered 496 and B) plant niche overlap and average number of seeds per fruit. Dots represent each of 497 the individuals sampled for each species within each site. 498 Figure 3. Partial residual plots showing the effect of A) pollinator richness, and B) niche 499 complementarity between pollinator species on site-level average fruitset and C) niche 500 complementarity between pollinator species on the average number of seeds per fruit at 501 the site level. Dots represent average values of fruit set at the level of the community for 502 all plant species considered (N=16 sites). 503 Figure 4. Partial residual plots showing the effect of pollinator richness on site-level 504 average A) fruit and B) seed weight. Dots represent values for each site (N=16 sites). 505 Figure 5. Results of simulation evaluating the importance of niche complementarity in 506 determining differences in equity in reproductive across communities harboring from one 507 to six species. Points represent average values across 1,000 simulated combinations.

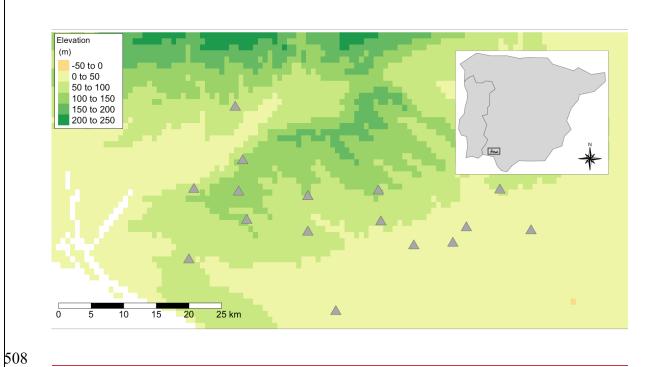
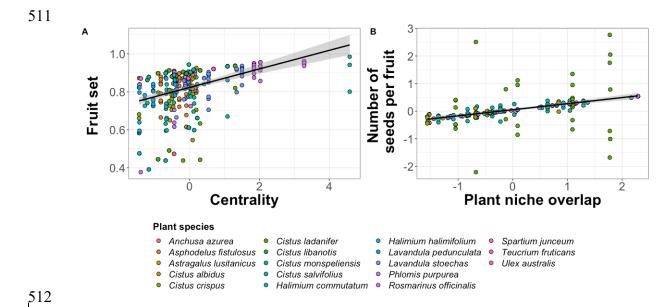


Figure 1.



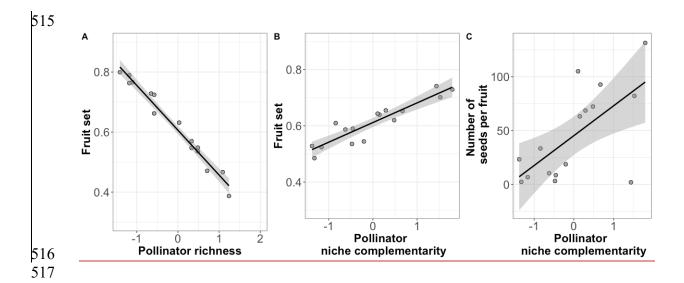
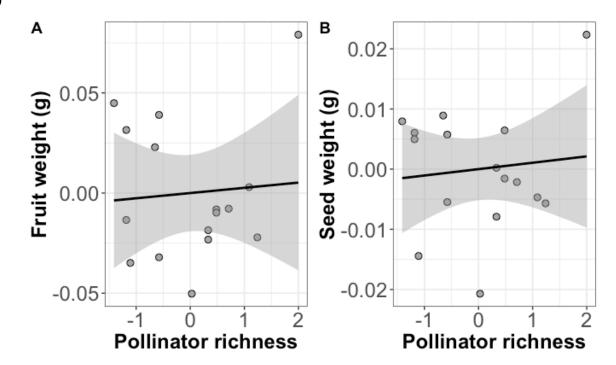
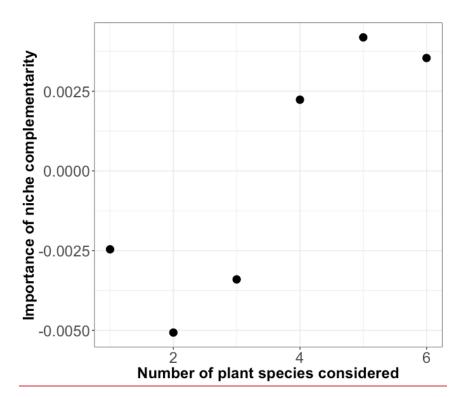


Figure 3.





522 Figure <u>4</u>.



525 <u>Figure 5.</u>

528 References

- Abrams, P.A., Holt, R.D. & Roth, J.D. (1998). Apparent Competition or Apparent
- Mutualism? Shared Predation When Populations Cycle. *Ecology*, 79, 201–212.
- Albrecht, M., Schmid, B., Hautier, Y. & Müller, C.B. (2012). Diverse pollinator
- communities enhance plant reproductive success. *Proceedings of the Royal Society B:*
- 533 *Biological Sciences*, 279, 4845 LP–4852.
- Almeida-Neto, M. & Ulrich, W. (2011). A straightforward computational approach for
- measuring nestedness using quantitative matrices. *Environ Model Softw*, 26, 173–178.
- Bagchi, R., Swinfield, T., Gallery, R.E., Lewis, O.T., Gripenberg, S. & Narayan, L. et al.
- 537 (2010). Testing the Janzen-Connell mechanism: pathogens cause overcompensating
- density dependence in a tropical tree. *Ecology Letters*, 13, 1262–1269.
- Bartomeus, I., Stavert, J., Ward, D. & Aguado, O. (2019). Historical collections as a tool
- for assessing the global pollination crisis. *Philosophical Transactions of the Royal*
- *Society B: Biological Sciences*, 374, 20170389.
- Bascompte, J. & Jordano, P. (2007). Plant-Animal Mutualistic Networks: The
- 543 Architecture of Biodiversity. *Annu Rev Ecol Evol S*, 38, 567–593.
- Bascompte, J., Jordano, P., Melián, C.J. & Olesen, J.M. (2003). The nested assembly of
- plant–animal mutualistic networks. P. Natl. Acad. Sci. USA, 100, 9383–9387.
- Bastolla, U., Fortuna, M.A., Pascual-Garcia, A., Ferrera, A., Luque, B. & Bascompte, J.
- 547 (2009). The architecture of mutualistic networks minimizes competition and increases
- 548 biodiversity. *Nature*, 458, 1018–1020.

- Benadi, G., Blüthgen, N., Hovestadt, T. & Poethke, H.-J. (2013). When can plant-
- pollinator interactions promote plant diversity? *Am Nat*, 182, 131–146.
- Benadi, G. & Pauw, A. (2018) Frequency dependence of pollination visitation rates
- suggests that pollination niches can allow plant species coexistence. Journal of Ecology,
- 553 106, 1892-1901.
- Bennett, J.M., Thompson, A., Goia, I., Feldmann, R., Ştefan, V. & Bogdan, A. et al.
- 555 (2018). A review of European studies on pollination networks and pollen limitation, and
- a case study designed to fill in a gap. *AoB PLANTS*, 10, ply068–ply068.
- 557 Biesmeijer, J.C., Roberts, S.P., Reemer, M., Ohlemuller, R., Edwards, M. & Peeters, T. et
- 558 al. (2006). Parallel declines in pollinators and insect-pollinated plants in Britain and the
- 559 Netherlands. *Science*, 313, 351–354.
- Blüthgen, N. (2010). Why network analysis is often disconnected from community
- ecology: A critique and an ecologist's guide. *Basic and Applied Ecology*, 11, 185–195.
- Blüthgen, N. & Klein, A.-M. (2011). Functional complementarity and specialisation: The
- role of biodiversity in plant–pollinator interactions. Basic and Applied Ecology, 12, 282–
- 564 <u>291.</u>
- Blüthgen, N., Menzel, F. & Blüthgen, N. (2006). Measuring specialization in species
- interaction networks. *BMC Ecology*, 6, 1–12.
- Brosi, B.J. (2016). Pollinator specialization: from the individual to the community. *New*
- 568 *Phytologist*, 210, 1190–1194.

- Bruno, J.F., Stachowicz, J.J. & Bertness, M.D. (2003). Inclusion of facilitation into
- ecological theory. *Trends in Ecology & Evolution*, 18, 119–125.
- Burnham, K.P., Anderson, D. & Huyvaert, K.P. (2011). AIC model selection and
- 572 multimodel inference in behavioral ecology: some background, observations, and
- 573 comparisons. Behav Ecol Sociobiol, 65, 23–35.
- 574 Carvalheiro, L.G., Biesmeijer, J.C., Benadi, G., Fründ, J., Stang, M. & Bartomeus, I. et
- 575 al. (2014). The potential for indirect effects between co-flowering plants via shared
- 576 pollinators depends on resource abundance, accessibility and relatedness. *Ecology*
- 577 Letters, 17, 1389–1399.
- 578 Chacoff, N.P., Vázquez, D.P., Lomáscolo, S.B., Stevani, E.L., Dorado, J. & Padrón, B.
- 579 (2012). Evaluating sampling completeness in a desert plant-pollinator network. *Journal*
- 580 *of Animal Ecology*, 81, 190–200.
- 581 Chao, A., Colwell, R.K., Lin, C. & Gotelli, N.J. (2009). Sufficient sampling for
- asymptotic minimum species richness estimators. *Ecology*, 90, 1125–1133.
- Delmas, E., Besson, M., Brice, M.-H., Burkle, L.A., Dalla Riva, G.V. & Fortin, M.-J. et
- 484 al. (2019). Analysing ecological networks of species interactions. *Biological Reviews*, 94,
- 585 <u>16–36.</u>
- 586 Devoto, M., Bailey, S., Craze, P. & Memmott, J. (2012). Understanding and planning
- 587 ecological restoration of plant–pollinator networks. *Ecology Letters*, 15, 319–328.
- Dormann, C.F., Frund, J., Bluthgen, N. & Gruber, B. (2009). Indices, graphs and null
- models: analyzing bipartite ecological networks. The Open Ecology Journal, 2, 7–24.

- Flanagan, R.J., Mitchell, R.J., Knutowski, D. & Karron, J.D. (2009). Interspecific
- 591 pollinator movements reduce pollen deposition and seed production in Mimulus ringens
- 592 (Phrymaceae). American Journal of Botany, 96, 809–815.
- 593 Fontaine, C., Dajoz, I., Meriguet, J. & Loreau, M. (2005). Functional Diversity of Plant-
- Pollinator Interaction Webs Enhances the Persistence of Plant Communities. *PLOS*
- 595 *Biology*, 4, e1.
- 596 Fründ, J., Dormann, C.F., Holzschuh, A. & Tscharntke, T. (2013). Bee diversity effects
- on pollination depend on functional complementarity and niche shifts. *Ecology*, 94,
- 598 2042–2054.
- 599 Garibaldi, L., Carvalheiro, L.G., Leonhardt, S.D., Aizen, M.A., Blaauw, B.R. & Isaacs,
- R. et al. (2015). From research to action: enhancing crop yield through wild pollinators.
- 601 Front Ecol Environ, 12, 439–447.
- 602 Garibaldi, L.A., Steffan-Dewenter, I., Winfree, R., Aizen, M.A., Bommarco, R. &
- 603 Cunningham, S.A. et al. (2013). Wild Pollinators Enhance Fruit Set of Crops Regardless
- of Honey Bee Abundance. *Science*, 339, 1608–1611.
- Godoy, O., Bartomeus, I., Rohr, R.P. & Saavedra, S. (2018). Towards the Integration of
- Niche and Network Theories. *Trends in Ecology & Evolution*, 33, 287–300.
- 607 Goldberg, D.E. & Barton, A.M. (1992). Patterns and Consequences of Interspecific
- 608 Competition in Natural Communities: A Review of Field Experiments with Plants. *The*
- 609 *American Naturalist*, 139, 771–801.

- 610 Holt, R.D. (1977). Predation, apparent competition, and the structure of prey
- 611 <u>communities. Theoretical Population Biology</u>, 12, 197–229.
- Hsieh, T.C., Ma, K.H. & Chao, A. (2016). iNEXT: iNterpolation and EXTrapolation for
- species diversity. R package version 2.0.8.
- James, A., Pitchford, J.W. & Plank, M.J. (2012). Disentangling nestedness from models
- of ecological complexity. *Nature*, 487, 227–230.
- Johnson, C.A. & Bronstein, J.L. (2019). Coexistence and competitive exclusion in
- 617 <u>mutualism. *Ecology*</u>, 100, e02708.
- Jordano, P. (2016) Sampling networks of ecological interactions. Functional ecology, 30,
- 619 1183-1893.
- 620 Lanuza, J.B., Bartomeus, I. & Godoy, O. (2018). Opposing effects of floral visitors and
- soil conditions on the determinants of competitive outcomes maintain species diversity in
- heterogeneous landscapes. *Ecology Letters*, 21, 865–874.
- 623 Lázaro, A., Lundgren, R. & Totland, Ø. (2014). Experimental reduction of pollinator
- visitation modifies plant–plant interactions for pollination. Oikos, 123, 1037–1048.
- Macarthur, R. & Levins, R. (1967). The Limiting Similarity, Convergence, and
- Divergence of Coexisting Species. *The American Naturalist*, 101, 377–385.
- May, R.M. & MacArthur, R.H. (1972). Niche Overlap as a Function of Environmental
- Variability. Proceedings of the National Academy of Sciences, 69, 1109 LP–1113.

- Mayfield, M.M. & Stouffer, D.B. (2017). Higher-order interactions capture unexplained
- 630 complexity in diverse communities. *Nature Ecology & Amp; Evolution*, 1, 62.
- Morales, C.L. & Traveset, A. (2008). Interspecific Pollen Transfer: Magnitude,
- 632 Prevalence and Consequences for Plant Fitness. Critical Reviews in Plant Sciences, 27,
- 633 221–238.
- Nakagawa, S., Johnson, P.C.D. & Schielzeth, H. (2017). The coefficient of determination
- R2 and intra-class correlation coefficient from generalized linear mixed-effects models
- revisited and expanded. *Journal of The Royal Society Interface*, 14, 20170213.
- Nieto, A., Roberts, S., Kemp, J., Rasmont, P., Kuhlmann, M. & García Criado, M. et al.
- 638 (2014). European Red List of Bees.
- Pauw, A. (2013) Can pollination niches facilitate plant coexistence? Trends in Ecology &
- 640 Evolution, 28, 30-37.
- Petchey, O.L. & Gaston, K.J. (2007). Dendrograms and measuring functional diversity.
- 642 *Oikos*, 116, 1422–1426.
- Poisot, T., Mouquet, N. & Gravel, D. (2013). Trophic complementarity drives the
- 644 biodiversity–ecosystem functioning relationship in food webs. *Ecology Letters*, 16, 853–
- 645 861.
- Potts, S.G., Biesmeijer, J.C., Kremen, C., Neumann, P., Schweiger, O. & Kunin, W.E.
- 647 (2010). Global pollinator declines: trends, impacts and drivers. *Trends Ecol Evol*, 25,
- 648 345–353.

- R Development Core Team, R. (2011). R: A Language and Environment for Statistical
- 650 Computing. R foundation for statistical computing.
- Simmons, B.I., Hoeppke, C. & Sutherland, W.J. (2019). Beware greedy algorithms.
- 652 *Journal of Animal Ecology*, 88, 804–807.
- Song, C., Rohr, R.P. & Saavedra, S. (2017). Why are some plant–pollinator networks
- more nested than others? *Journal of Animal Ecology*, 86, 1417–1424.
- Stavert, J.R., Bartomeus, I., Beggs, J.R., Gaskett, A.C. & Pattemore, D.E. (2019). Plant
- species dominance increases pollination complementarity and plant reproductive
- 657 <u>function</u>. *Ecology*, 0, e02749.
- Thébault, E. & Fontaine, C. (2010). Stability of Ecological Communities and the
- Architecture of Mutualistic and Trophic Networks. *Science*, 329, 853–856.
- Thompson, R.M., Brose, U., Dunne, J.A., Hall Jr., R.O., Hladyz, S. & Kitching, R.L. et
- al. (2012). Food webs: reconciling the structure and function of biodiversity. Trends in
- 662 Ecology & Evolution, 27, 689–697.
- Thomson, D.M. (2019). Effects of long-term variation in pollinator abundance and
- diversity on reproduction of a generalist plant. *Journal of Ecology*, 0.
- Valdovinos, F.S., Brosi, B.J., Briggs, H.M., Moisset de Espanés, P., Ramos-Jiliberto, R.
- & Martinez, N.D. (2016). Niche partitioning due to adaptive foraging reverses effects of
- 667 nestedness and connectance on pollination network stability. *Ecology Letters*, 19, 1277–
- 668 1286.

- Vanbergen, A., Woodcock, B., Gray, A., Grant, F., Telford, A., Lambdon, P., Chapman,
- D.S., Pywell, R.F., Heard, M.S. & Cavers, S. (2014) Grazing alters insect visitation
- networks and plant mating systems. Functional Ecology, 28, 178-189.
- Winfree, R. (2013). Global change, biodiversity, and ecosystem services: What can we
- learn from studies of pollination? *Basic and Applied Ecology*, 14, 453–460.
- 674 Woodward, G. & Hildrew, A.G. (2002). Body-size determinants of niche overlap and
- 675 <u>intraguild predation within a complex food web. Journal of Animal Ecology</u>, 71, 1063–
- 676 1074.
- 677 Zhang, J. (2016). Package spaa: SPecies Association Analysis.